

1 Biting Diptera-host network structure varies 2 with anthropogenic landscape modification

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23 Abstract

- 24 1. Rapid and ongoing anthropogenic habitat modification has the potential to alter the
25 species composition, abundance and activity of biting insect communities, which are
26 important disease vectors. The resulting changes in the network of interactions
27 between biting insects and their hosts have implications for the transmission of
28 vector-borne pathogens.
- 29 2. We used DNA metabarcoding of Diptera blood meals to document bipartite networks
30 of interactions between biting flies (Diptera) and their hosts (including humans,
31 domesticated and wild animals) across a gradient of anthropogenic habitat
32 modification (village, agricultural and near-natural habitat) surrounding two rural
33 villages in Ghana.
- 34 3. We collected 7,095 biting Diptera (of 42 species) from 30 collection sites, and
35 generated sequencing data from 75 blood meals (from 29 species). These blood meals
36 contained DNA from 18 vertebrate host species, dominated by humans and their
37 livestock.
- 38 4. Habitats with lower levels of anthropogenic modification had higher richness of biting
39 Diptera and their host species. Species diversity and evenness did not differ
40 significantly among habitats. Less modified habitats had higher network specificity,
41 but connectance was highest in heavily modified habitats.
- 42 5. Humans were highly embedded within biting Diptera-host networks, detected in 68%
43 of blood meals. The networks reveal several potential disease transmission pathways
44 linking competent vectors with susceptible hosts. The presence of mixed blood meals

45 containing DNA of both human and wild animal origin highlights the potential for
46 transmission of established and emerging zoonotic disease via bridge vectors. The
47 high betweenness-centrality within interaction networks of the important disease
48 vector *Culex watti*, combined with its high abundance across all levels of
49 anthropogenic landscape modification, suggest that it may be a connector species,
50 linking and facilitating disease transmission between spatially distinct communities.

51 6. *Synthesis and applications*: Our results are of epidemiological interest, as they
52 identify the exposure of humans to pathogen transmission cycles across a gradient of
53 anthropogenic habitat modification through the movement of opportunistic bridge
54 vectors. We discuss the implications for the transmission of emerging and established
55 zoonotic disease and for the targeting and implementation of initiatives to reduce
56 disease exposure and transmission.

57 Keywords

58 Anthropogenic landscape modification, biting Diptera, disease transmission, ecological
59 networks

60 Introduction

61 Landscape modification characteristic of the Anthropocene can directly and indirectly drive
62 changes in communities at the boundary between modified and natural habitats, increasing
63 the risk of pathogen transmission. Biting Diptera, feeding as adults on the blood of
64 vertebrates, transmit a wide range of pathogens that affect humans, as well as both wild and
65 domestic animals, and represent a significant threat to public and economic wellbeing
66 (Braack et al., 2018). Globally, an estimated one-sixth of all illness and disability is linked to

67 vector-borne diseases, with over half of the global population currently at risk of infection
68 (Campbell-Lendrum et al., 2015). Low- and Middle-income countries (LMIC) in the Global
69 South are disproportionately at risk, with over 831 million people (70% of the population) at
70 risk of infection from vector borne diseases in Africa alone (Agboli et al., 2021; Benelli and
71 Beier, 2017). These risks are linked to extensive on-going changes in natural landscapes for
72 agricultural and urban expansion (Chaves et al., 2020; Jones et al., 2013).

73 Anthropogenic landscape modification increases contact between humans, domestic and wild
74 animals, and associated vectors, impacting epidemiological processes (Gibb et al., 2020b;
75 Mayi et al., 2019; Patz et al., 2004). Rural populations that rely on subsistence farming are
76 especially at risk, because vector abundances are positively associated with the primary
77 sources of rural income (especially livestock) (Franklinos et al., 2019). Rural populations are
78 also often situated adjacent to sylvatic habitats, allowing pathogens to cycle between wild
79 animals and vectors, increasing exposure to established and emerging zoonoses (Gibb et al.,
80 2020a). Anthropogenic land-use change fragments the structure of the landscape, producing
81 smaller patches of sylvatic habitat surrounded by agricultural or urban land (Chaves et al.,
82 2010). The resulting increase in edge habitats increases interaction frequencies and increases
83 the potential for opportunistic biting Diptera species to act as bridge vectors (Hoyos et al.,
84 2021; Meyer Steiger et al., 2016). Furthermore, host species diversity declines as human
85 land-use intensifies (Abella-Medrano et al., 2018). This can increase disease transmission
86 potential by eroding the dilution effect (where high host diversity limits the probability of a
87 vector encountering a competent host (Civitello et al., 2015; Miller and Huppert, 2013)).

88 Consequently, habitats with limited host diversity (e.g. urban or agricultural land) likely have
89 higher zoonotic transmission and spillover, especially if these habitats are dominated by
90 competent intermediate or ‘amplifier’ hosts, within which a pathogen may replicate rapidly
91 and be maintained at a high concentration (Franklinos et al., 2019; Jones et al., 2013).

92 Landscape modification can also alter the diversity and composition of biting Diptera
93 communities by changing aspects of the biotic and abiotic environment, such as the
94 availability of standing water, vegetation cover and host availability (Chaves et al., 2021;
95 Ferraguti et al., 2016; Li et al., 2014). Depending on their particular lifecycles and habitat
96 requirements, different biting Diptera species are likely to vary in their responses to such
97 landscape changes (Perrin et al., 2022). For example, for species with aquatic larvae,
98 increased local temperatures and artificial water bodies providing breeding habitats outside
99 the normal seasonal rainfall periods in agricultural and village habitats decreases larval
100 development time and allows reproduction independent of seasonal precipitation cycles
101 (Gratz, 1999; Kesavaraju et al., 2008; Mattah et al., 2017). Human land-use is thus expected
102 to alter and homogenise assemblages of biting Diptera and their vertebrate hosts, increasing
103 transmission and spillover potential and altering networks of trophic interactions (Chaves et
104 al., 2021; Ferraguti et al., 2016; Van Hoesel et al., 2019).

105 Despite the intimate and important interconnections among humans, livestock, biting Diptera
106 and pathogens (Bellekom et al., 2021), empirical interaction data allowing for quantitative
107 network analyses remains limited, particularly in West Africa. Here we investigate
108 interactions of biting Diptera with their hosts in two rural villages in Ghana. Increasing
109 anthropogenic landscape modification, expansion in agricultural land, and human
110 encroachment into sylvatic habitats in recent decades (Food and Agriculture Organization of
111 the United Nations, 2023; Acheampong et al., 2019) mean that a significant proportion of the
112 Ghanaian population now live and work at the interface between rural and sylvatic habitat,
113 with implications for the community structure of biting Diptera, their associated host
114 interactions, and disease transmission potential.

115 We use PacBio Sequel sequencing of Diptera bloodmeals to investigate how networks linking
116 biting Diptera (including known or potential disease vectors) to their vertebrate hosts differed

117 between anthropogenically-modified habitats. We test the hypothesis that anthropogenic
118 landscape modification would homogenise the community of vectors and hosts, altering the
119 structure of bipartite interaction networks, and increasing the centrality of humans within
120 them.

121 Materials and Methods

122 Study sites

123 We sampled insects around two rural villages within the Volta region, Ghana: Abutia
124 Amageme (AA) (6.459064°N, 0.316247°E) and Mafi Agove (MA) (6.208310°N 0.442600°E;
125 **Error! Reference source not found.**). Both are small, farming-dependent villages with
126 cultivated land primarily used for maize and cassava surrounded by a semi-natural mixture of
127 forest and grassland. Abutia Amegame retains more natural and semi-natural vegetation than
128 Mafi Agove, which is larger with a less heterogeneous landscape matrix; see (Hemprich-
129 Bennett et al., 2026) for full site descriptions. No biting Diptera control initiatives were in
130 place during the sampling period, although some residents sleep under insecticide-treated bed
131 nets.

132 Insect collection

133 Insect sampling at each village took place between July 2019 and February 2020, over a 24-
134 hour period every three weeks. On each visit, we sampled at four locations adjacent to each
135 village. Each location was positioned randomly each week within each of four quadrants
136 (NE, NW, SE, SW) of a 500 m radius circle centred on the village. We captured biting
137 Diptera using CDC (Centre for Disease Control and Prevention) light traps (Bioquip, Rancho
138 Dominguez, CA) 1.3 metres above the ground and modified to include a collection chamber

139 filled with 95% ethanol (Bellekom et al., 2023; **Error! Reference source not found.**), with
140 one trap at each of the four sampling locations. We used a sugar fermented yeast CO₂ bait,
141 consisting of 1L of water, 100g of sugar and 7g of yeast within a 1.5L bottle, attached to the
142 trap entrance (Supplementary Fig. 2; Jerry et al., 2017; Smallegange et al., 2010).
143 Anthropogenic habitat modification at each sampling location was assessed using a
144 combination of field-based assessment of vegetation type and Google Earth satellite imagery.
145 Locations where cultivated land or livestock were the dominant land-use were categorised as
146 Agricultural; those containing natural vegetation with limited human presence were classified
147 as Near-natural; and sampling locations in or around human habitation were classified as
148 Village. Data were accumulated from 30 collection sites: 12 Agricultural, 12 Near-natural,
149 and six Village.

150 Sample processing

151 Insect samples were stored soon after collection and for up to 48-hours in a refrigerator and
152 then transported in cold boxes to a freezer at the University of Ghana, Accra. DNA
153 amplification success from bloodmeals stored in 95% ethanol is not strongly affected by
154 storage temperature (Bellekom et al., 2023), however, time at room temperature was still kept
155 to a minimum. Prior to molecular identification, samples were visually sorted to Order and
156 assigned a morphospecies. Up to five representatives of each morphospecies were placed in
157 96-well microplates containing 95% ethanol, and the total abundance of each morphospecies
158 per sampling event was recorded. Blood-fed individuals (those with a visible blood meal) and
159 potentially blood-fed individuals (gravid and parous individuals, and those with almost
160 entirely digested blood meals) were visually identified and separated into a different 96-well
161 plate for blood meal analysis. All plated samples were shipped under permit in cold boxes to
162 the Centre for Biodiversity Genomics facility at the University of Guelph for DNA
163 sequencing.

164 Blood meal analysis

165 Samples were processed using automated pipelines established by the Canadian Centre for
166 DNA Barcoding. Diptera and blood meal DNA were handled by a Biomek FXP liquid
167 handling system and extracted using a glass fibre extraction protocol (Ivanova et al., 2006).
168 Single molecule real-time (SMRT) sequencing was carried out on the PacBio Sequel II
169 platform, targeting the COI gene. Amplicons were generated using the SMRTbell Express
170 Template kit 2.0, Sequel II Binding kit 2.1 and Sequel II DNA Internal Control 1.0 (Hebert et
171 al., 2018).

172 Bioinformatics

173 Sequencing data were initially processed using the mBRAVE platform. Sequences were
174 quality filtered and clustered into OTUs using default parameters, with a 2% sequence
175 divergence threshold. Low-abundance OTUs (<2% of reads per sample) were excluded.
176 Chimeric sequences, false reads, and controls were manually removed. A custom COI
177 reference library was created using NSDPY (Hebert and Megléc, 2022) and blastn (Wheeler
178 and Bhagwat, 2007), incorporating families expected in the region based on a manual search
179 of GBIF (gbif.org, 2021) and field guides (Grubb, 1998); (Borrow and Demey, 2010).
180 Taxonomic assignment used blastn (word size = 28, max_target_seqs = 1) and a 98%
181 sequence identity threshold.
182 As individual samples may generate multiple contigs, singleton contigs within mixed samples
183 were removed as likely artefacts. Any singletons observed only once in the entire dataset
184 were retained, as they were deemed to likely be true results rather than contamination. Where
185 a sample was represented by multiple contigs with differing species assignments, contigs
186 were retained only if they were not singletons of expected fragment length (~180bp) and
187 ≥98% identity match (Hopken et al., 2021). Where a sample could not be identified to species

188 level, the highest taxonomic resolution was assigned. Samples that retained contigs of
189 different species origin after filtering were classified as having a mixed blood meal. To
190 minimise potential human contamination, contigs identified as human DNA were only
191 retained if they contained >10 reads in a sample, based on observed read distributions of
192 confirmed non-human host taxa (e.g. *Philantomba maxwellii*).
193 Blood-fed and non-blood-fed Diptera were processed using the same mBRAVE protocol.
194 Where taxonomy could not be assigned automatically using mBRAVE, sequences were
195 manually identified using BLAST. Where a sample could not be identified to species level, a
196 unique identifier was assigned to prevent multiple species being represented by a single
197 family identification in downstream analysis. Abundances were assigned to each occurrence
198 of a species using the estimated abundance for each morphospecies. Where a morphospecies
199 was represented by multiple species, the abundance of each component species was assigned
200 based on their overall relative proportion.

201 Data analysis

202 We explored the effect of habitat type on log-transformed Diptera Shannon-Wiener diversity
203 using a generalized linear model with a Gaussian error distribution. The statistical
204 significance of habitat type was assessed by comparing nested models for variation in
205 deviance based on a chi-square distribution (Mayi et al., 2020). Post hoc analysis was
206 conducted to identify differences between individual habitat types using a Tukey's HSD
207 (honestly significant difference) test. The dissimilarity of the biting Diptera community
208 composition across habitat types and locations was investigated using a Permutational
209 Multivariate Analysis of Variance Using Distance Matrices (PERMANOVA) (Oksanen et al.,
210 2025).
211 Interaction data were pooled across sampling locations by habitat type to generate a weighted
212 bipartite network for each category. To examine the complexity of our networks, we

213 calculated their connectance, the fraction of potential ecological interactions that are realised
214 (Poisot and Gravel, 2014). Network specialisation for each habitat classification was
215 examined using H_2' , which quantifies the deviation of observed interaction frequencies from
216 the frequencies expected if interactions are random (Blüthgen et al., 2006). As generalist
217 species are more likely to act as ‘connector species’ (Hackett et al., 2019) that promote cross-
218 species disease transmission, networks with a high degree of generality may sustain a greater
219 disease transmission potential between unrelated hosts (Su et al., 2022). To assess the risk of
220 disease exposure potential for humans in each habitat type, we calculated the degree and
221 closeness centrality for humans in each network. As the within-species prevalence of a
222 disease increases linearly with its degree, nodes with a high degree (i.e., many connections to
223 the target node) are more likely to participate in disease transmission events (Su et al., 2022).
224 Closeness centrality describes the mean path length from the target node to all other nodes in
225 the network, with a high value indicating that the node may be rapidly affected by other
226 nodes in the network and vice versa (Silk et al., 2017; Su et al., 2022). To examine the
227 relative importance of a node in connecting parts of each network and their potential to
228 mediate transmission throughout the network, we calculated the betweenness centrality of the
229 biting Diptera in each of our networks. Betweenness centrality describes the proportion of
230 shortest pathlengths in a network that go via the node, with a high value indicating that the
231 node is important for the cohesiveness of the network (Llaberia-Robledillo et al., 2022).
232 While we did not directly identify pathogens within the blood meals, we inferred disease
233 vectoring potential of each blood-fed Diptera species using Vectorbase (GiraldoCalderón et
234 al., 2015) to highlight potential transmission events for susceptible hosts. To assess sampling
235 completeness, we constructed interpolation and extrapolation curves for hosts, biting Diptera
236 and their interactions as a function of sampling effort (the number of blood meals collected)
237 using the iNEXT package ($q=0$, data type = incidence frequency) (Hsieh et al., 2016).

238 Finally, we estimated total interactions, and host and Diptera species richness using the
239 ChaoRichness function (Hsieh et al., 2016), which is based on methods proposed by (Chao,
240 1987).

241 Results

242 Overview

243 In total we collected 7,095 biting Diptera from the 30 collection sites (12 Agricultural, 12
244 Near-natural, and six Village sites) (**Error! Reference source not found., Error! Reference**
245 **source not found.**). We identified 42 individual biting Diptera species (84% of nodes) and
246 resolved an additional eight biting Diptera to genus or family level (16% of nodes) (**Error!**
247 **Reference source not found.**1). Blood-fed individuals were collected from 20 sampling
248 locations: 6 Agricultural, 11 Near-natural, and 3 Village. Blood meal sequencing on the
249 PacBio Sequel II platform generated an initial 189,240 reads. After filtering and removal of
250 chimeras this was reduced to 184,938 reads. Approximately half (95,772) of our reads were
251 assigned to a BIN, whilst 89,166 were placed in 11,016 OTUs. Following progression
252 through our BLAST pipeline, using our custom reference library for host species
253 identification, this was reduced to 111,449 reads. After implementing our strict filtering
254 criteria, we retained 3,151 reads which characterised the composition of 75 blood meals and
255 identified 18 individual host species derived from blood meals of 29 biting Diptera species.

256

257 Table 1: The proportion of fed individuals for each biting Diptera species sampled across all
258 sites.

Species	Total	Blood-fed	Proportion blood-fed
<i>Culex nebulosus</i>	4	1	25.00%
<i>Culex watti</i>	175	24	13.71%
<i>Aedes bromeliae</i>	2	1	50.00%

<i>Culex theileri</i>	2	0	0.00%
<i>Aedeomyia africana</i>	2	0	0.00%
<i>Culicoides imicola</i>	1649	2	0.12%
<i>Culicoides hortensis</i>	130	3	2.31%
<i>Sergentomyia ingrami</i>	17	1	5.88%
<i>Sergentomyia hamoni</i>	18	1	5.56%
<i>Sergentomyia africana</i>	227	1	0.44%
<i>Culex rima</i>	9	3	33.33%
<i>Mimomyia mimomyiformis</i>	12	7	58.33%
<i>Coquillettidia metallica</i>	20	2	10.00%
<i>Culicoides distinctipennis</i>	2766	2	0.07%
<i>Mimomyia mediolineata</i>	4	0	0.00%
<i>Culex perexiguus</i>	8	2	25.00%
<i>Sergentomyia inermis</i>	337	1	0.30%
<i>Culex aurantapex</i>	2	0	0.00%
<i>Mansonia africana</i>	25	8	32.00%
<i>Culex decens</i>	2	1	50.00%
<i>Forcipomyia sp. 1</i>	347	0	0.00%
<i>Culicoides oxystoma</i>	130	0	0.00%
<i>Sergentomyia schwetzi</i>	40	1	2.50%
<i>Sergentomyia sp. 1</i>	41	1	2.44%
<i>Dasyhelea ludingensis</i>	555	0	0.00%
<i>Sergentomyia clydei</i>	15	0	0.00%
<i>Culex univittatus</i>	7	1	14.29%
<i>Psychoda sp. 1</i>	140	0	0.00%
<i>Culex tritaeniorhynchus</i>	7	1	14.29%
<i>Culex cinereus</i>	1	1	100.00%
<i>Aedes unilineatus</i>	3	0	0.00%
<i>Uranotaenia alboabdominalis</i>	3	0	0.00%
<i>Aedes hirsutus</i>	4	0	0.00%
<i>Culex duttoni</i>	2	0	0.00%
<i>Culex naevei</i>	3	0	0.00%
<i>Anopheles coustani</i>	4	0	0.00%
<i>Culex vansomerini</i>	4	1	25.00%
<i>Culex adersianus</i>	2	0	0.00%
<i>Aedes quasiunivittatus</i>	2	0	0.00%
<i>Culicoides neavei</i>	3	0	0.00%
<i>Uranotaenia bilineata</i>	2	0	0.00%
<i>Culex annulioris</i>	2	0	0.00%
<i>Culicoides enderleini</i>	257	0	0.00%

<i>Ceratopogonidae sp.1</i>	4	1	25.00%
<i>Sergentomyia distincta</i>	12	3	25.00%
<i>Ceratopogonidae sp.2</i>	45	1	2.22%
<i>Ceratopogonidae sp.3</i>	42	1	2.38%
<i>Uranotaenia micromelas</i>	3	1	33.33%
<i>Atrichopogon sp.1</i>	4	1	25.00%
<i>Allohelea sp.1</i>	1	1	100.00%

259

260 Table 1: Non-blood-fed and blood-fed Diptera richness, number of blood meals, and host
261 richness by level of anthropogenic landscape modification.

Site/Habitat	Non-blood-fed Diptera richness	Non-blood-fed Diptera abundance	Blood-fed Diptera richness	Number of blood meals	Host richness
Village	11	1,382	4	8	2
Agricultural	29	1,454	11	28	9
Near-natural	34	4,259	22	39	11

262 Effect of habitat modification

263 Non-blood-fed and blood-fed Diptera richness and abundance was lower in Abutia Amegame
264 than in Mafi Agove (**Error! Reference source not found.2**). Total richness and abundance
265 of non-blood-fed and blood-fed Diptera were lower in the Village habitats compared to
266 Agricultural and Near-natural habitat, as was host richness (Table 1: The proportion of fed
267 individuals for each biting Diptera species sampled across all sites.

Species	Total	Blood-fed	Proportion blood-fed
<i>Culex nebulosus</i>	4	1	25.00%
<i>Culex watti</i>	175	24	13.71%
<i>Aedes bromeliae</i>	2	1	50.00%
<i>Culex theileri</i>	2	0	0.00%
<i>Aedeomyia africana</i>	2	0	0.00%
<i>Culicoides imicola</i>	1649	2	0.12%
<i>Culicoides hortensis</i>	130	3	2.31%

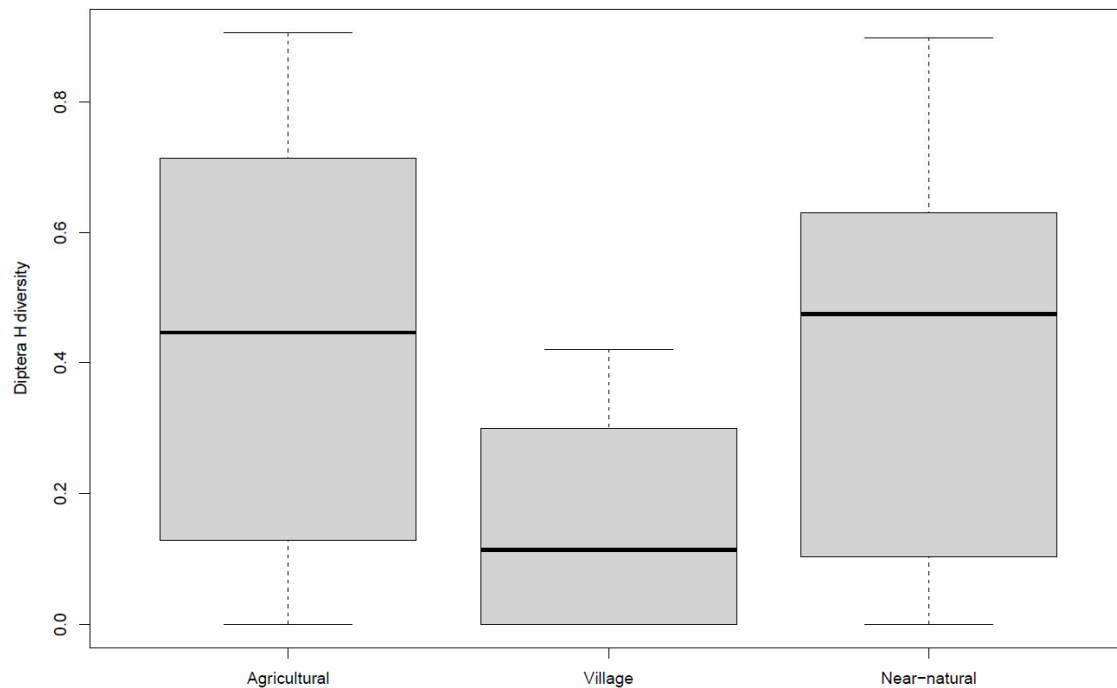
<i>Sergentomyia ingrami</i>	17	1	5.88%
<i>Sergentomyia hamoni</i>	18	1	5.56%
<i>Sergentomyia africana</i>	227	1	0.44%
<i>Culex rima</i>	9	3	33.33%
<i>Mimomyia mimomyiformis</i>	12	7	58.33%
<i>Coquillettidia metallica</i>	20	2	10.00%
<i>Culicoides distinctipennis</i>	2766	2	0.07%
<i>Mimomyia mediolineata</i>	4	0	0.00%
<i>Culex perexiguus</i>	8	2	25.00%
<i>Sergentomyia inermis</i>	337	1	0.30%
<i>Culex aurantapex</i>	2	0	0.00%
<i>Mansonia africana</i>	25	8	32.00%
<i>Culex decens</i>	2	1	50.00%
<i>Forcipomyia sp. 1</i>	347	0	0.00%
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<i>Sergentomyia sp. 1</i>	41	1	2.44%
<i>Dasyhelea ludingensis</i>	555	0	0.00%
<i>Sergentomyia clydei</i>	15	0	0.00%
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<i>Culex cinereus</i>	1	1	100.00%
<i>Aedes unilineatus</i>	3	0	0.00%
<i>Uranotaenia alboabdominalis</i>	3	0	0.00%
<i>Aedes hirsutus</i>	4	0	0.00%
<i>Culex duttoni</i>	2	0	0.00%
<i>Culex naevei</i>	3	0	0.00%
<i>Anopheles coustani</i>	4	0	0.00%
<i>Culex vansomerini</i>	4	1	25.00%
<i>Culex adersianus</i>	2	0	0.00%
<i>Aedes quasiunivittatus</i>	2	0	0.00%
<i>Culicoides neavei</i>	3	0	0.00%
<i>Uranotaenia bilineata</i>	2	0	0.00%
<i>Culex annulioris</i>	2	0	0.00%
<i>Culicoides enderleini</i>	257	0	0.00%
<i>Ceratopogonidae sp.1</i>	4	1	25.00%
<i>Sergentomyia distincta</i>	12	3	25.00%
<i>Ceratopogonidae sp.2</i>	45	1	2.22%
<i>Ceratopogonidae sp.3</i>	42	1	2.38%

<i>Uranotaenia micromelas</i>	3	1	33.33%
<i>Atrichopogon sp.1</i>	4	1	25.00%
<i>Allohelea sp.1</i>	1	1	100.00%

268

269 Table 1). Biting Diptera species evenness was similar ($F_{2,20} = 1.570$; $p = 0.233$) in Village
270 (mean = 0.109, SE = 0.04), Agricultural (mean = 0.272, SE = 0.06) and Near-natural (mean =
271 0.220, SE 0.04) habitats. There was no significant effect of habitat modification on biting
272 Diptera diversity ($\chi^2 = 0.29178$; $p = 0.186$) (Figure 1).

273

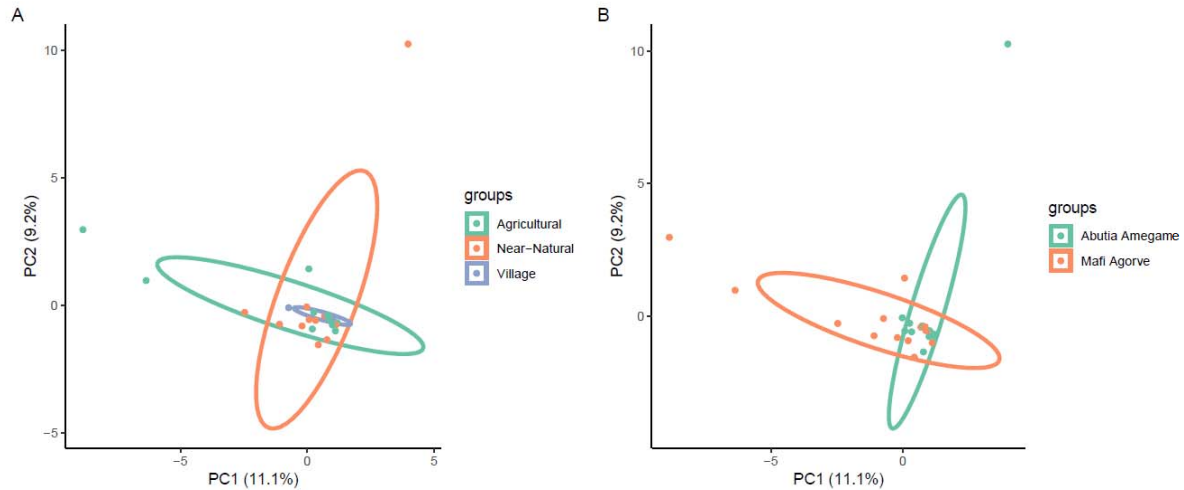


274

275 Figure 1: Biting Diptera Shannon (H) diversity by habitat type. Each box displays the
276 interquartile range, and the solid line represents the median. Whiskers display the maximum
277 and minimum interaction evenness and H values for each habitat type.

278 Biting Diptera community composition did not differ significantly across habitat types ($F_{2,27}$
279 = 1.09; $p = 0.306$; Figure 2A) or site ($F_{1,28} = 0.862$; $p = 0.611$; **Error! Reference source not
280 found.**).

281



282

283 Figure 2: Ordination of the overall biting Diptera community composition by A) habitat
284 category or B) Village. Each point represents the biting Diptera diversity captured by a
285 unique sampling event in Village (purple) Agricultural (blue), and Near-natural (orange)
286 habitat. Or in Abutia Amegame (blue) and Mafi Agorve (orange).

287

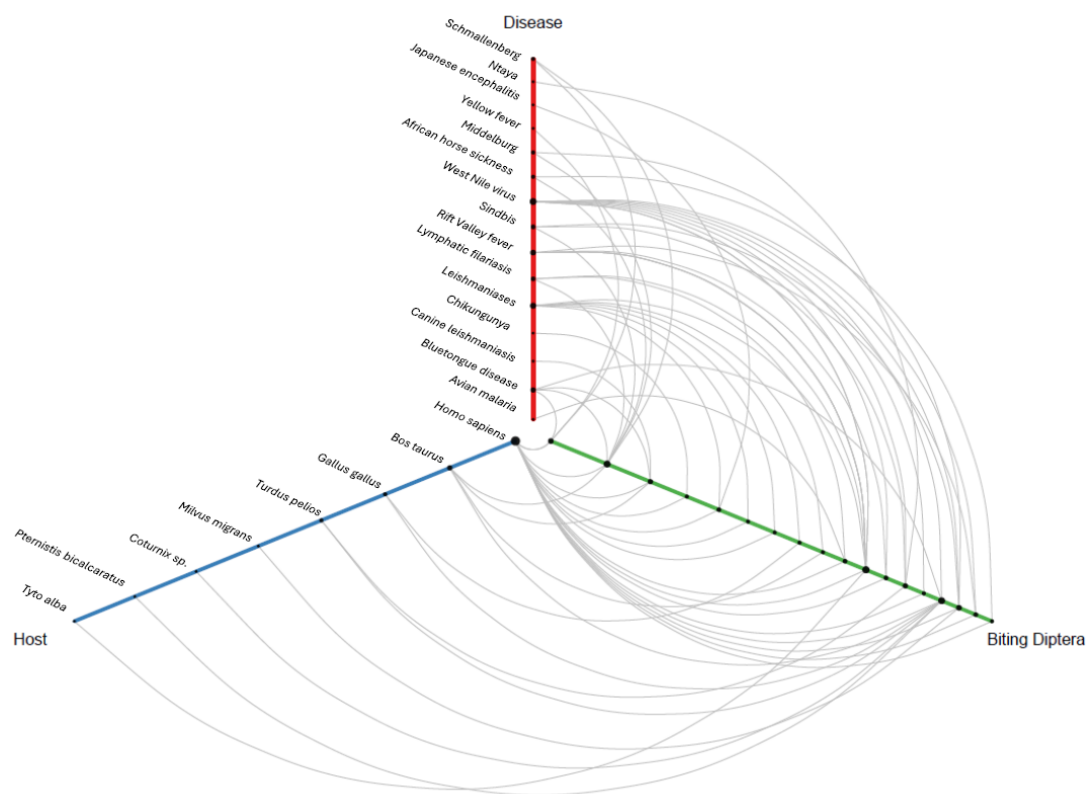
288 Blood meal origin

289 Humans had the highest number of interactions across all sites and habitat classifications;
290 human DNA was detected in 68% of all collected blood meals. The number of wild host
291 species differed markedly between the villages, with 10 species in Mafi Agorve and just one
292 in Abutia Amegame, with minimal difference in domestic animal species richness (3 and 4
293 species, respectively). There were slightly more domestic and wild host species in Near-
294 natural habitat (4 and 6) than in Agricultural habitat (3 and 5). In Village habitats we only
295 identified interactions with human, and one mixed blood meal containing human DNA and
296 DNA from a wild bird species (a quail, *Coturnix* sp.).
297 Mixed blood meals accounted for 7.89% (6) of all blood meals. Three mixed blood meals
298 were found in samples from Near-natural habitat (7.69%), two from Agricultural habitat
299 (7.14%) and one from Village (12.5%). Five of the six mixed blood meals contained human

300 DNA. Two of the mixed blood meals contained DNA from three host origins (Soricidae,
301 Fringillidae, and Phasianinae) and (*Capra hircus*, *Ovis aries*, and *Homo sapiens*).

302 Potential disease transmission events

303 Of the 29 blood-fed biting Diptera species collected, 18 (62%) were competent of vectoring
304 one or more pathogen of medical or veterinary importance. Of the vector-competent Diptera
305 species, 16 (88%) had interacted with at least one host species susceptible to the pathogens
306 they vector (Figure 3; Supplementary Table 3).



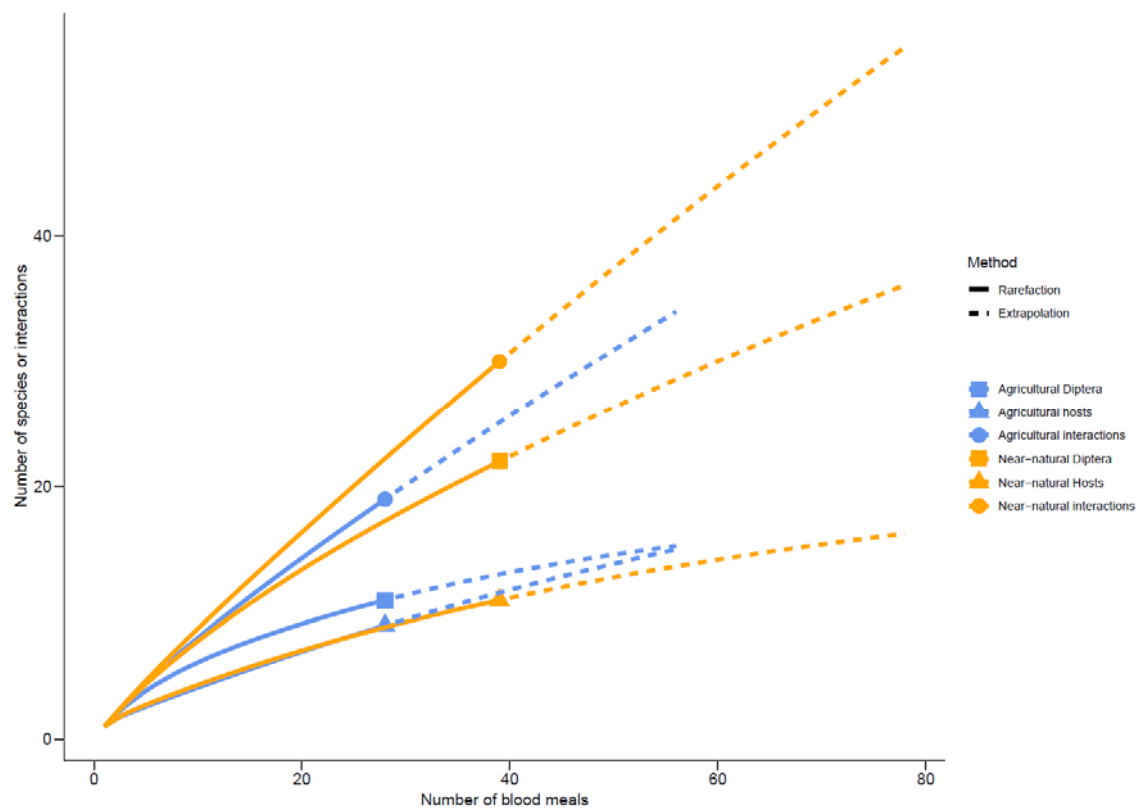
307

308 Figure 3: Hive plot showing how potential transmission events between hosts (blue), biting
309 Diptera (green) and their respective diseases (red). Node size along each axis scale with
310 frequency. Diptera nodes, from inside to outside, are: *Culex perexiguus*, *Culicoides imicola*,
311 *Culicoides distinctipennis*, *Sergentomyia schwetzi*, *Aedes bromeliae*, *Sergentomyia ingrami*,
312 *Sergentomyia Africana*, *Sergentomyia hamoni*, *Sergentomyia inermis*,
313 *Sergentomyia distincta*, *Mansonia Africana*, *Culex decens*, *Culex univittatus*, *Culex cinereus*,
314 *Culex watti*, *Coquillettidia metallica*, *Culex nebulosus*, *Culex vansomereni*.

315 Community structure

316 Accumulation curves showed that host and biting Diptera species richness for Agricultural
317 and host richness for Near-natural habitat were relatively well resolved, with curves
318 approaching an asymptote in each case. Sampling for biting Diptera species in Near-natural
319 habitat and species interactions in all habitats were incomplete (Figure 44 and **Error!**
320 **Reference source not found.**). Due to the lack of interaction data, Village habitats were
321 excluded from sampling completeness analysis.

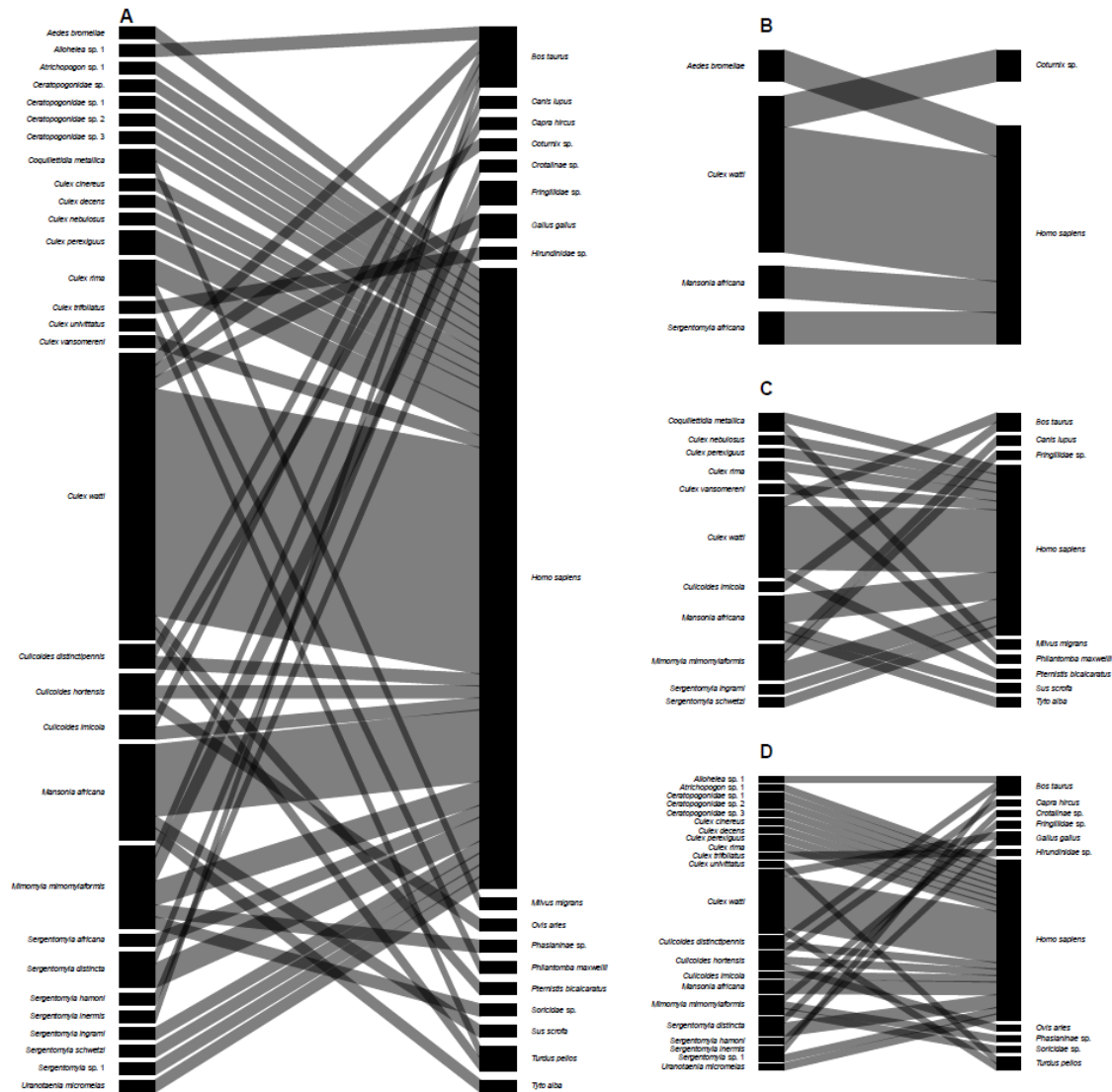
322 All interaction data were compiled into an overall summary network (Figure A), and separate
323 networks were also constructed for each habitat category (Fig. 5B-D) and each location
324 (Supplementary Fig.4). The connectance of the overall network was 0.088. Whilst no
325 statistical comparison of network metrics across habitats was possible due to lack of
326 replication, the Village network had higher connectance (0.625) than either the Agricultural
327 (0.191) or the Near-natural (0.123) network, although sample sizes in terms of numbers of
328 blood meals were very low for the Village sites.



329

330 Figure 4: Smoothed accumulation and extrapolation curves to assess sampling completeness.
331 Total numbers of hosts (triangle), biting Diptera (square), and interactions recorded in the
332 whole dataset (circle), by habitat type (Agricultural in blue, Near-natural in orange), as a
333 function of sampling effort (the number of blood meals screened). Village habitat was
334 excluded from this analysis due to under sampling.

335



336

337 Figure 5: Combined habitat networks. (A) The aggregated network containing all host (right)
 338 and Diptera (left) interactions, interactions separated by habitat classification: Village (B),
 339 Agricultural (C), and Near-natural (D). Node and edge widths are proportional to frequency
 340 of occurrence.

341 The H2' specialisation of the combined network across both sites was 0.192, and network
 342 specialisation was higher in Near-natural (0.347) than Agricultural (0.148) habitats. There
 343 were insufficient data to reliably calculate the Village H2' value.

344 Human closeness centrality was 0.094 and degree 25. Closeness centrality was higher in
 345 Village habitat (0.5) than Agricultural (0.170) and Near-natural habitat (0.171). The degree of

346 humans was higher in Near-natural habitat (16) than in Agricultural (10) and Village habitats
347 (4).

348 *Culex watti* had the highest betweenness centrality of all the blood-fed biting Diptera in our
349 combined network (0.450). Moreover, *C. watti* had the highest betweenness centrality across
350 both Near-natural (0.507) and Agricultural (1) habitats. Due to the lack of data, Village was
351 excluded from this analysis.

352 Discussion

353 Our results revealed variations in the richness of biting Diptera across a gradient of
354 anthropogenic habitat modification, with higher richness and abundance in Near-natural and
355 Agricultural habitats. Network metrics also differed with anthropogenic landscape
356 modification, and the detailed data on individual vector species and their feeding interactions
357 that emerge from our networks reveal a wide variety of potential transmission pathways
358 between biting Diptera and disease-competent hosts.

359 Ecological networks populated with interaction data generated through the DNA sequencing
360 of biting Diptera blood meals provide an insight into the community and interaction structure
361 of Diptera and their hosts in a sampled location. Whilst every effort was taken to reduce bias,
362 there are several potential limitations to this approach. Interaction networks provide a
363 snapshot of community and interaction structure, and multiple years of sampling are required
364 to document temporal changes in network structure (Olivier et al., 2019). To minimise the
365 effect of varying diel activity patterns, for each sampling event we sampled over a period of
366 24 hours. Under-sampling may alter conclusions drawn from several network metrics
367 (Blüthgen, 2010); the rarefaction analysis demonstrated that interactions, and host and
368 Diptera richness, are indeed under-sampled. Significantly increased sampling effort, beyond

369 what was possible within time and travel limitations, would be required to increase the
370 confidence in the network metrics values (Ings et al., 2009)

371 Overall abundance of biting Diptera decreased with landscape modification, with a biting
372 Diptera abundance nearly three times higher in Near-natural habitat compared to other
373 habitats. This could reflect the proximity and availability of breeding sites and hosts
374 (Barrientos-Roldán et al., 2022; Ferraguti et al., 2016; Johnston et al., 2014; Pereira-Silva et
375 al., 2021; Young et al., 2021). Indeed, despite highly modified land, such as Village habitat,
376 providing breeding opportunities for biting Diptera outside of seasonal precipitation cycles,
377 only those synanthropic species capable of propagating and feeding on hosts in these habitats
378 are likely to be present (Barata et al., 2012; Johnston et al., 2014). Consequently, the higher
379 biting Diptera species richness and abundance in Near-natural and Agricultural habitat are
380 likely due to the habitat providing a greater number of hosts and more variable breeding sites
381 that support more species and higher abundances of each species. The observed pattern of
382 increased host richness within blood meals in Near-natural habitat is consistent with the
383 expectation that Near-natural habitats retain more wild species. This host diversity could help
384 reduce disease transmission in less modified habitat through the dilution effect (Shah et al.,
385 2019).

386 Species evenness was low across all habitats, but was lowest in village interiors, consistent
387 with previous studies (Möhlmann et al., 2017; Petruff et al., 2020; Visa Shalini et al., 2022).

388 Several species were abundant across all habitat types, including *Culicoides imicola*,
389 *Culicoides distinctipennis*, *Sergentomyia africana* and *Culex watti*. These species are likely
390 habitat generalists with a degree of plasticity in blood feeding behaviour, allowing them to
391 exploit a wide range of hosts (Stone and Gross, 2018; Villard and Metzger, 2014). Indeed, we
392 recorded *Culex watti* as having blood meals from six host species. As these abundant species
393 are competent vectors of a range of pathogens, their abundance and wide distribution have

394 significant implications for the emergence of zoonotic diseases and disease transmission
395 potential across a range of habitats.

396 We found no significant dissimilarity in community composition across habitats. However,
397 we identified biting Diptera species that were unique to each habitat. Indeed, 53% of Diptera
398 species collected in the Near-natural classification were unique to this habitat, perhaps
399 reflecting a restriction to habitat conditions absent in human-modified parts of the landscape
400 (Vittor et al., 2009). Communities in transitional zones (ecotones) often comprise species
401 from the adjacent habitats (Fortin et al., 2000). We observed a pattern of high Diptera species
402 overlap across an anthropogenic landscape modification gradient from Agricultural to Near-
403 natural (34% of Agricultural species were also present in Near-natural habitat), indicating a
404 shared community and consistent with movement between the two habitat types (Meyer
405 Steiger et al., 2016). Overlapping biting Diptera communities link Agricultural with adjacent
406 Near-natural habitat and, consequently, humans with the sylvatic cycle through the
407 movement of opportunistic bridge vectors (Miot et al., 2020). We found several species that
408 link the Village and Agricultural habitat, including *Anopheles coustani*. This species is a
409 known vector of diseases including malaria, and displays anthropophilic behaviour
410 (Ciubotariu et al., 2020). These findings are epidemiologically relevant as they indicate that
411 existing and future Agricultural habitat may facilitate disease transmission across the
412 landscape and provide a pathway through which emerging and established zoonotic disease
413 are introduced into rural and urban transmission cycles.

414 Network connectance was greater within the Village habitat than Agricultural and Near-
415 natural habitats, although this may reflect the limited number of recorded interactions and
416 interaction partners, and the abundance of interactions with humans (Valdovinos et al., 2009).
417 Low network connectance in Agricultural and Near-natural habitat is likely a result of the
418 increased host species richness (Rivera-Hutinel et al., 2012). Interpolation and extrapolation

419 curves demonstrated that biting Diptera-host interactions were highly under-sampled in
420 comparison to richness estimates, in all habitat types. Consequently, connectance values in
421 our study may not be representative of the true interaction structure (Blüthgen, 2010; Heleno
422 et al., 2012). Whilst most networks are incomplete (Vizentin & Bugoni et al., 2016),
423 incomplete sampling, as in the current study, can still be sufficient to capture the majority of
424 functionally important species in a network (Hegland et al., 2010).

425 Network specialisation (H2') decreased with anthropogenic landscape modification, although
426 values were uniformly low across habitats, likely reflecting plasticity in host use. Host usage
427 has a high degree of plasticity in biting Diptera and may be strongly influenced by local host
428 densities (Takken and Verhulst, 2013). Increased H2' in Near-natural habitat could reflect the
429 greater host and Diptera density, species richness, as well as potential niche partitioning
430 (Chakravarty et al., 2023; Dalsgaard et al., 2011). Lower network specialism in Agricultural
431 than Near-natural habitat is particularly epidemiologically relevant, as generalist vectors have
432 an increased capacity to transmit emerging and established zoonoses into novel hosts
433 (Santiago-Alarcon et al., 2012). Therefore, habitat with a lower network specialism will
434 likely have greater disease transmission potential (Ellwanger and Chies, 2021). Indeed,
435 agricultural drivers have previously been linked to over 50% of all zoonotic disease
436 emergence (Rohr et al., 2019). Whilst each habitat network was considered distinct and
437 sampling conducted over an extended period provided a snapshot of community structure in
438 each habitat classification, the contiguous nature of the landscape, and overlapping Diptera
439 communities, suggest that movement of individuals between habitats is likely. Such
440 movement has the potential to link susceptible hosts in one network with competent vectors
441 in another, facilitating disease transmission across the landscape.

442 The high degree centrality of humans in Agricultural and Near-natural habitat demonstrates
443 that humans may act as a hub species (Toju et al., 2018), interacting with a wide range of

444 biting Diptera species (including disease-competent vectors) and (indirectly) other hosts.
445 Indeed, the presence of mixed blood meals containing DNA of human and wild animal origin
446 highlights the potential for transmission of established and emerging zoonotic disease via
447 bridge vectors (Brackney et al., 2021; Miot et al., 2020).
448 The closeness centrality of humans was greatest in Village habitat and there was minimal
449 difference between Agricultural and Near-natural habitat, suggesting that transmission
450 potential to humans is higher in anthropogenically modified habitat (Grubb et al., 2021;
451 Ribeiro et al., 2020). These findings are particularly epidemiologically relevant, as they
452 suggest that, at least in the study landscape, humans may be equally at risk of contracting
453 vector-borne diseases when working in Agricultural land as they are when encroaching or
454 traversing Near-natural habitat. The high centrality of humans in Village habitats is expected
455 due to the anthropophilic feeding behaviour of the sampled blood-fed Diptera and their
456 historical association with anthropogenic habitat (Bennett et al., 2015; Ughasi et al., 2012).
457 However, high centrality may also reflect the limited number of blood meals collected from
458 Village locations (Costenbader and Valente, 2003). Indeed, given the observed presence of
459 domestic hosts, such as goats, chickens, and dogs, there was a surprising lack of domestic
460 animal host DNA in the Village blood meals. This could in part reflect anthropophilic feeding
461 behaviour, but is perhaps more likely a consequence of under-sampling of areas with high
462 densities of dwellings, which were restricted to the centre of the two sampling areas. Despite
463 lower closeness centrality values in Agricultural and Near-natural habitat, high human degree
464 centrality in these habitats remains an epidemiological concern (Bell et al., 1999).
465 Across the range of blood-fed biting Diptera collected in all habitat classifications, *Culex*
466 *watti*, a potential vector of West Nile Virus (Diarra et al., 2019), consistently had the highest
467 betweenness centrality. A high betweenness centrality value indicates that the species is
468 important for the cohesion of the network and favours the circulation of their vectored

469 pathogens throughout the network (Espinaze et al., 2018; Llaberia-Robledillo et al., 2022).
470 Whilst a higher number of blood-fed and non-blood-fed *C. watti* were collected in
471 Agricultural and Near-natural habitats, the presence of *C. watti* within all habitat
472 classifications indicates a degree of generality in their habitat range (Stone and Gross, 2018).
473 Moreover, the high betweenness values and presence across all levels of anthropogenic
474 landscape modifications suggest that *C. watti* has the potential to act as a connector species,
475 linking and facilitating disease transmission between spatially distinct communities
476 (Bellekom et al., 2021; Llaberia-Robledillo et al., 2022).
477 Increasing anthropogenic landscape modification could drive a range of vector-borne disease
478 outbreaks and zoonotic emergence events through its effect on biting Diptera and host
479 community composition and vector-host relationships (Meyer Steiger et al., 2016; Patz et al.,
480 2004). The characterisation of the biting Diptera-host community and interaction structure
481 provides opportunities to highlight and implement control initiatives to mitigate transmission
482 potential. In the short term, such opportunities may include behavioural modification for
483 village inhabitants to limit exposure to transmission, including informing locals of potential
484 risk factors, limiting encroachment into Near-natural habitat, covering the skin and using
485 insect repellent when working at the interface between Agricultural and Near-natural habitat
486 to minimise biting, and increase uptake of insecticide-treated bed netting. Further, in the
487 medium term, such interaction data may highlight vectors and hosts of particular importance
488 to network cohesion and disease circulation. These data will inform localised control
489 initiatives, for example, targeting synanthropic breeding habitat to suppress anthropophilic
490 species, and the identification and removal of key domestic host species that may act as
491 reservoirs of disease. In the long term, Diptera blood meal networks may be used to inform
492 potential ecological consequences and network rewiring that may occur following vector
493 control initiatives.

494 The high degree of overlap in community composition and shared host usage between
495 Agricultural and Near-natural land places rural populations that rely on subsistence farming
496 at the interface of disease transmission between rural and sylvatic habitat (Meyer Steiger et
497 al., 2016). Closeness and degree centrality metrics indicated that there was a high disease
498 transmission risk to humans. Incorporation of complementary quantitative susceptibility and
499 vector competency rates may provide a more resolved assessment of vector-human
500 transmission potential. Our interaction data, when combined with vector competency data,
501 indicated a high number of potential transmission pathways. Subsequent interaction and
502 pathogen data would thus allow for network approaches to be applied on an individual basis,
503 in which nodes represent host or Diptera individuals and edges potential transmission events
504 (Bellekom et al., 2021). Such data would be highly epidemiologically relevant and allow for
505 the monitoring of the transmission of established and emerging zoonosis across the rural-
506 sylvatic interface.

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519 Author contributions statement

520 BB, OTL and TDH conceived the ideas and designed methodology; BB, NAA, BA, ED and
521 TDH collected the data with support from FAA; BB analysed the data with support from
522 DRHB, TDH and OTL; OTL, FAA and TDH provided project supervision; BB led the
523 writing of the manuscript. All authors contributed critically to the drafts and gave final
524 approval for publication.

525

526 Data archiving statement

527 Data is archived on the Zenodo digital repository: <https://doi.org/10.5281/zenodo.19591599>

528 Conflicts of interest

529 None listed

530 Data availability statement

531 Data available from the Zenodo digital repository: <https://doi.org/10.5281/zenodo.19591599>

532 Code used to run the analysis and generate figures is available on [github](#)

533 References

- 534 Abella-Medrano, C.A., Ibáñez-Bernal, S., Carbó-Ramírez, P., Santiago-Alarcon, D., 2018.
535 Blood-meal preferences and avian malaria detection in mosquitoes (Diptera:
536 Culicidae) captured at different land use types within a neotropical montane cloud
537 forest matrix. *Parasitology International* 67, 313–320.
538 <https://doi.org/10.1016/j.parint.2018.01.006>
- 539 Acheampong, E.O., Macgregor, C.J., Sloan, S., Sayer, J., 2019. Deforestation is driven by
540 agricultural expansion in Ghana’s forest reserves. *Scientific African* 5, e00146.
541 <https://doi.org/10.1016/j.sciaf.2019.e00146>
- 542 Agboli, E., Zahouli, J.B.Z., Badolo, A., Jöst, H., 2021. Mosquito-Associated Viruses and
543 Their Related Mosquitoes in West Africa. *Viruses* 13, 891.
544 <https://doi.org/10.3390/v13050891>
- 545 Barata, R.A., Ursine, R.L., Nunes, F.P., Morais, D.H., Araújo, H.S., 2012. Synanthropy of
546 mosquitoes and sand flies near the Aimorés hydroelectric power plant, Brazil. *Journal*
547 *of Vector Ecology* 37, 397–401. <https://doi.org/10.1111/j.1948-7134.2012.00243.x>
- 548 Barrientos-Roldán, M.J., Abella-Medrano, C.A., Ibáñez-Bernal, S., Sandoval-Ruiz, C.A.,
549 2022. Landscape Anthropization Affects Mosquito Diversity in a Deciduous Forest in
550 Southeastern Mexico. *Journal of Medical Entomology* 59, 248–256.
551 <https://doi.org/10.1093/jme/tjab154>
- 552 Bellekom, B., Bailey, A., England, M., Langlands, Z., Lewis, O.T., Hackett, T.D., 2023.
553 Effects of storage conditions and digestion time on DNA amplification of biting
554 midge (Culicoides) blood meals. *Parasites Vectors* 16, 13.
555 <https://doi.org/10.1186/s13071-022-05607-x>

- 556 Bellekom, B., Hackett, T.D., Lewis, O.T., 2021. A Network Perspective on the Vectoring of
557 Human Disease. *Trends in Parasitology* 37, 391–400.
558 <https://doi.org/10.1016/j.pt.2020.12.001>
- 559 Benelli, G., Beier, J.C., 2017. Current vector control challenges in the fight against malaria.
560 *Acta Tropica* 174, 91–96. <https://doi.org/10.1016/j.actatropica.2017.06.028>
- 561 Blüthgen, N., 2010. Why network analysis is often disconnected from community ecology: A
562 critique and an ecologist’s guide. *Basic and Applied Ecology* 11, 185–195.
563 <https://doi.org/10.1016/j.baae.2010.01.001>
- 564 Blüthgen, Nico, Menzel, F., Blüthgen, Nils, 2006. Measuring specialization in species
565 interaction networks. *BMC Ecology* 6. <https://doi.org/10.1186/1472-6785-6-9>
- 566 Borrow, N., Demey, R., 2010. *Birds of Ghana, Helm field guides*. C. Helm, London.
- 567 Braack, L., Gouveia de Almeida, A.P., Cornel, A.J., Swanepoel, R., de Jager, C., 2018.
568 Mosquito-borne arboviruses of African origin: review of key viruses and vectors.
569 *Parasites Vectors* 11, 29. <https://doi.org/10.1186/s13071-017-2559-9>
- 570 Brackney, D.E., LaReau, J.C., Smith, R.C., 2021. Frequency matters: How successive
571 feeding episodes by blood-feeding insect vectors influences disease transmission.
572 *PLoS Pathog* 17, e1009590. <https://doi.org/10.1371/journal.ppat.1009590>
- 573 Campbell-Lendrum, D., Manga, L., Bagayoko, M., Sommerfeld, J., 2015. Climate change
574 and vector-borne diseases: what are the implications for public health research and
575 policy? *Phil. Trans. R. Soc. B* 370, 20130552. <https://doi.org/10.1098/rstb.2013.0552>
- 576 Chakravarty, R., Radchuk, V., Managave, S., Voigt, C.C., 2023. Increasing species richness
577 along elevational gradients is associated with niche packing in bat assemblages.
578 *Journal of Animal Ecology* 92, 863–874. <https://doi.org/10.1111/1365-2656.13897>
- 579 Chao, A., 1987. Estimating the Population Size for Capture-Recapture Data with Unequal
580 Catchability. *Biometrics* 43, 783. <https://doi.org/10.2307/2531532>

- 581 Chaves, L.F., Harrington, L.C., Keogh, C.L., Nguyen, A.M., Kitron, U.D., 2010. Blood
582 feeding patterns of mosquitoes: random or structured? *Front Zool* 7, 3.
583 <https://doi.org/10.1186/1742-9994-7-3>
- 584 Chaves, L.S.M., Bergo, E.S., Conn, J.E., Laporta, G.Z., Prist, P.R., Sallum, M.A.M., 2021.
585 Anthropogenic landscape decreases mosquito biodiversity and drives malaria vector
586 proliferation in the Amazon rainforest. *PLoS ONE* 16, e0245087.
587 <https://doi.org/10.1371/journal.pone.0245087>
- 588 Chaves, L.S.M., Fry, J., Malik, A., Geschke, A., Sallum, M.A.M., Lenzen, M., 2020. Global
589 consumption and international trade in deforestation-associated commodities could
590 influence malaria risk. *Nat Commun* 11, 1258. [https://doi.org/10.1038/s41467-020-](https://doi.org/10.1038/s41467-020-14954-1)
591 [14954-1](https://doi.org/10.1038/s41467-020-14954-1)
- 592 Ciobotariu, I.I., Jones, C.M., Kobayashi, T., Bobanga, T., Muleba, M., Pringle, J.C.,
593 Stevenson, J.C., Carpi, G., Norris, D.E., 2020. Genetic Diversity of *Anopheles*
594 *coustani* (Diptera: Culicidae) in Malaria Transmission Foci in Southern and Central
595 Africa. *Journal of Medical Entomology* 57, 1782–1792.
596 <https://doi.org/10.1093/jme/tjaa132>
- 597 Civitello, D.J., Cohen, J., Fatima, H., Halstead, N.T., Liriano, J., McMahon, T.A., Ortega,
598 C.N., Sauer, E.L., Sehgal, T., Young, S., Rohr, J.R., 2015. Biodiversity inhibits
599 parasites: Broad evidence for the dilution effect. *Proc. Natl. Acad. Sci. U.S.A.* 112,
600 8667–8671. <https://doi.org/10.1073/pnas.1506279112>
- 601 Costenbader, E., Valente, T.W., 2003. The stability of centrality measures when networks are
602 sampled. *Social Networks* 25, 283–307. [https://doi.org/10.1016/S0378-](https://doi.org/10.1016/S0378-8733(03)00012-1)
603 [8733\(03\)00012-1](https://doi.org/10.1016/S0378-8733(03)00012-1)
- 604 Dalsgaard, B., Magård, E., Fjeldså, J., Martín González, A.M., Rahbek, C., Olesen, J.M.,
605 Ollerton, J., Alarcón, R., Cardoso Araujo, A., Cotton, P.A., Lara, C., Machado, C.G.,

606 Sazima, I., Sazima, M., Timmermann, A., Watts, S., Sandel, B., Sutherland, W.J.,
607 Svenning, J.-C., 2011. Specialization in Plant-Hummingbird Networks Is Associated
608 with Species Richness, Contemporary Precipitation and Quaternary Climate-Change
609 Velocity. PLoS ONE 6, e25891. <https://doi.org/10.1371/journal.pone.0025891>

610 Diarra, A.Z., Laroche, M., Berger, F., Parola, P., 2019. Use of MALDI-TOF MS for the
611 Identification of Chad Mosquitoes and the Origin of Their Blood Meal. The American
612 Journal of Tropical Medicine and Hygiene 100, 47–53.
613 <https://doi.org/10.4269/ajtmh.18-0657>

614 Ellwanger, J.H., Chies, J.A.B., 2021. Zoonotic spillover: Understanding basic aspects for
615 better prevention. Genet. Mol. Biol. 44, e20200355. [https://doi.org/10.1590/1678-](https://doi.org/10.1590/1678-4685-gmb-2020-0355)
616 [4685-gmb-2020-0355](https://doi.org/10.1590/1678-4685-gmb-2020-0355)

617 Espinaze, M.P.A., Hellard, E., Horak, I.G., Cumming, G.S., 2018. Domestic mammals
618 facilitate tick-borne pathogen transmission networks in South African wildlife.
619 Biological Conservation 221, 228–236. <https://doi.org/10.1016/j.biocon.2018.03.017>

620 Ferraguti, M., Martínez-de La Puente, J., Roiz, D., Ruiz, S., Soriguer, R., Figuerola, J., 2016.
621 Effects of landscape anthropization on mosquito community composition and
622 abundance. Sci Rep 6, 29002. <https://doi.org/10.1038/srep29002>

623 Fortin, M.-J., Olson, R.J., Ferson, S., Iverson, L., Hunsaker, C., Edwards, G., Levine, D.,
624 Butera, K., Klemas, V., 2000. Issues related to the detection of boundaries. Landscape
625 Ecology 15, 453–466. <https://doi.org/10.1023/A:1008194205292>

626 Franklins, L.H.V., Jones, K.E., Redding, D.W., Abubakar, I., 2019. The effect of global
627 change on mosquito-borne disease. The Lancet Infectious Diseases 19, e302–e312.
628 [https://doi.org/10.1016/s1473-3099\(19\)30161-6](https://doi.org/10.1016/s1473-3099(19)30161-6)

629 gbif.org, 2021. Global Biodiversity Information Facility [WWW Document]. URL
630 <https://www.gbif.org>

- 631 Gibb, R., Franklinos, L.H.V., Redding, D.W., Jones, K.E., 2020a. Ecosystem perspectives are
632 needed to manage zoonotic risks in a changing climate. *BMJ* m3389.
633 <https://doi.org/10.1136/bmj.m3389>
- 634 Gibb, R., Redding, D.W., Chin, K.Q., Donnelly, C.A., Blackburn, T.M., Newbold, T., Jones,
635 K.E., 2020b. Zoonotic host diversity increases in human-dominated ecosystems.
636 *Nature* 584, 398–402. <https://doi.org/10.1038/s41586-020-2562-8>
- 637 Gratz, N.G., 1999. Emerging and resurging vector-borne diseases. *Annu. Rev. Entomol.* 44,
638 51–75. <https://doi.org/10.1146/annurev.ento.44.1.51>
- 639 Grubb, J., Lopez, D., Mohan, B., Matta, J., 2021. Network centrality for the identification of
640 biomarkers in respondent-driven sampling datasets. *PLoS ONE* 16, e0256601.
641 <https://doi.org/10.1371/journal.pone.0256601>
- 642 Grubb, P. (Ed.), 1998. *Mammals of Ghana, Sierra Leone and The Gambia*. The Trendrine
643 Press, St. Ives.
- 644 Hackett, T.D., Sauve, A.M.C., Davies, N., Montoya, D., Tylianakis, J.M., Memmott, J., 2019.
645 Reshaping our understanding of species' roles in landscape-scale networks. *Ecol Lett*
646 22, 1367–1377. <https://doi.org/10.1111/ele.13292>
- 647 Hebert, P.D.N., Braukmann, T.W.A., Prosser, S.W.J., Ratnasingham, S., deWaard, J.R.,
648 Ivanova, N.V., Janzen, D.H., Hallwachs, W., Naik, S., Sones, J.E., Zakharov, E.V.,
649 2018. A Sequel to Sanger: amplicon sequencing that scales. *BMC Genomics* 19, 219.
650 <https://doi.org/10.1186/s12864-018-4611-3>
- 651 Hebert, R., Megléc, E., 2022. NSDPY: A python package to download DNA sequences
652 from NCBI. *SoftwareX* 18, 101038. <https://doi.org/10.1016/j.softx.2022.101038>
- 653 Hegland, S.J., Dunne, J., Nielsen, A., Memmott, J., 2010. How to monitor ecological
654 communities cost-efficiently: The example of plant–pollinator networks. *Biological*
655 *Conservation* 143, 2092–2101. <https://doi.org/10.1016/j.biocon.2010.05.018>

- 656 Heleno, R., Devoto, M., Pocock, M., 2012. Connectance of species interaction networks and
657 conservation value: Is it any good to be well connected? *Ecological Indicators* 14, 7–
658 10. <https://doi.org/10.1016/j.ecolind.2011.06.032>
- 659 Hemprich-Bennett, D.R., Donkor, E., Adams, B., Acquah, N.A., Ofori, E.D., Anie-Amoah,
660 S., Bailey, A., Godfray, H.C.J., Lewis, O.T., Aboagye-Antwi, F., Hackett, T.D., 2026.
661 Characterising a species-rich and understudied tropical insect fauna using DNA
662 barcoding. *GigaScience* giag028. <https://doi.org/10.1093/gigascience/giag028>
- 663 Hopken, M.W., Reyes-Torres, L.J., Scavo, N., Piaggio, A.J., Abdo, Z., Taylor, D., Pierce, J.,
664 Yee, D.A., 2021. Temporal and Spatial Blood Feeding Patterns of Urban Mosquitoes
665 in the San Juan Metropolitan Area, Puerto Rico. *Insects* 12, 129.
666 <https://doi.org/10.3390/insects12020129>
- 667 Hoyos, J., Carrasquilla, M.C., León, C., Montgomery, J.M., Salyer, S.J., Komar, N.,
668 González, C., 2021. Host selection pattern and flavivirus screening of mosquitoes in a
669 disturbed Colombian rainforest. *Sci Rep* 11, 18656. [https://doi.org/10.1038/s41598-](https://doi.org/10.1038/s41598-021-98076-8)
670 [021-98076-8](https://doi.org/10.1038/s41598-021-98076-8)
- 671 Hsieh, T.C., Ma, K.H., Chao, A., 2016. iNEXT: an R package for rarefaction and
672 extrapolation of species diversity (Hill numbers). *Methods in Ecology and Evolution*
673 7, 1451–1456. <https://doi.org/10.1111/2041-210X.12613>
- 674 Ings, T.C., Montoya, J.M., Bascompte, J., Blüthgen, N., Brown, L., Dormann, C.F., Edwards,
675 F., Figueroa, D., Jacob, U., Jones, J.I., Lauridsen, R.B., Ledger, M.E., Lewis, H.M.,
676 Olesen, J.M., Van Veen, F.J.F., Warren, P.H., Woodward, G., 2009. Review:
677 Ecological networks – beyond food webs. *Journal of Animal Ecology* 78, 253–269.
678 <https://doi.org/10.1111/j.1365-2656.2008.01460.x>

- 679 Ivanova, N.V., Dewaard, J.R., Hebert, P.D.N., 2006. An inexpensive, automation-friendly
680 protocol for recovering high-quality DNA. *Molecular Ecology Notes* 6, 998–1002.
681 <https://doi.org/10.1111/j.1471-8286.2006.01428.x>
- 682 Jerry, D.C.T., Mohammed, T., Mohammed, A., 2017. Yeast-generated CO₂: A convenient
683 source of carbon dioxide for mosquito trapping using the BG-Sentinel® traps. *Asian*
684 *Pacific Journal of Tropical Biomedicine* 7, 896–900.
685 <https://doi.org/10.1016/j.apjtb.2017.09.014>
- 686 Johnston, E., Weinstein, P., Slaney, D., Flies, A.S., Fricker, S., Williams, C., 2014. Mosquito
687 communities with trap height and urban-rural gradient in Adelaide, South Australia:
688 implications for disease vector surveillance. *Journal of Vector Ecology* 39, 48–55.
689 <https://doi.org/10.1111/j.1948-7134.2014.12069.x>
- 690 Jones, B.A., Grace, D., Kock, R., Alonso, S., Rushton, J., Said, M.Y., McKeever, D., Mutua,
691 F., Young, J., McDermott, J., Pfeiffer, D.U., 2013. Zoonosis emergence linked to
692 agricultural intensification and environmental change. *Proc. Natl. Acad. Sci. U.S.A.*
693 110, 8399–8404. <https://doi.org/10.1073/pnas.1208059110>
- 694 Kesavaraju, B., Damal, K., Juliano, S.A., 2008. Do natural container habitats impede invader
695 dominance? Predator-mediated coexistence of invasive and native container-dwelling
696 mosquitoes. *Oecologia* 155, 631–639. <https://doi.org/10.1007/s00442-007-0935-4>
- 697 Li, Y., Kamara, F., Zhou, G., Puthiyakunnon, S., Li, C., Liu, Y., Zhou, Y., Yao, L., Yan, G.,
698 Chen, X.-G., 2014. Urbanization Increases *Aedes albopictus* Larval Habitats and
699 Accelerates Mosquito Development and Survivorship. *PLoS Negl Trop Dis* 8, e3301.
700 <https://doi.org/10.1371/journal.pntd.0003301>
- 701 Llaberia-Robledillo, M., Balbuena, J.A., Sarabeev, V., Llopis-Belenguer, C., 2022. Changes
702 in native and introduced host–parasite networks. *Biol Invasions* 24, 543–555.
703 <https://doi.org/10.1007/s10530-021-02657-7>

- 704 Mattah, P.A.D., Futagbi, G., Amekudzi, L.K., Mattah, M.M., De Souza, D.K., Kartey-
705 Attipoe, W.D., Bimi, L., Wilson, M.D., 2017. Diversity in breeding sites and
706 distribution of Anopheles mosquitoes in selected urban areas of southern Ghana.
707 Parasites Vectors 10, 25. <https://doi.org/10.1186/s13071-016-1941-3>
- 708 Mayi, M.P.A., Bamou, R., Djiappi-Tchamen, B., Fontaine, A., Jeffries, C.L., Walker, T.,
709 Antonio-Nkondjio, C., Cornel, A.J., Tchuinkam, T., 2020. Habitat and Seasonality
710 Affect Mosquito Community Composition in the West Region of Cameroon. Insects
711 11, 312. <https://doi.org/10.3390/insects11050312>
- 712 Mayi, M.P.A., Foncha, D.F., Kowo, C., Tchuinkam, T., Brisco, K., Anong, D.N., Ravinder,
713 S., Cornel, A.J., 2019. Impact of deforestation on the abundance, diversity, and
714 richness of *Culex* mosquitoes in a southwest Cameroon tropical rainforest. Journal of
715 Vector Ecology 44, 271–281. <https://doi.org/10.1111/jvec.12359>
- 716 Meyer Steiger, D.B., Ritchie, S.A., Laurance, S.G.W., 2016. Mosquito communities and
717 disease risk influenced by land use change and seasonality in the Australian tropics.
718 Parasites Vectors 9, 387. <https://doi.org/10.1186/s13071-016-1675-2>
- 719 Miller, E., Huppert, A., 2013. The Effects of Host Diversity on Vector-Borne Disease: The
720 Conditions under Which Diversity Will Amplify or Dilute the Disease Risk. PLoS
721 ONE 8, e80279. <https://doi.org/10.1371/journal.pone.0080279>
- 722 Miot, E.F., Calvez, E., Aubry, F., Dabo, S., Grandadam, M., Marcombe, S., Oke, C., Logan,
723 J.G., Brey, P.T., Lambrechts, L., 2020. Risk of arbovirus emergence via bridge
724 vectors: case study of the sylvatic mosquito *Aedes malayensis* in the Nakai district,
725 Laos. Sci Rep 10, 7750. <https://doi.org/10.1038/s41598-020-64696-9>
- 726 Möhlmann, T.W.R., Wennergren, U., Tälle, M., Favia, G., Damiani, C., Bracchetti, L.,
727 Koenraadt, C.J.M., 2017. Community analysis of the abundance and diversity of

728 mosquito species (Diptera: Culicidae) in three European countries at different
729 latitudes. *Parasites Vectors* 10, 510. <https://doi.org/10.1186/s13071-017-2481-1>

730 Oksanen, J., Simpson, G.L., Blanchet, F.G., Kindt, R., Legendre, P., Minchin, P.R., O'Hara,
731 R.B., Solymos, P., Stevens, M.H.H., Szoecs, E., Wagner, H., Barbour, M., Bedward,
732 M., Bolker, B., Borcard, D., Borman, T., Carvalho, G., Chirico, M., De Caceres, M.,
733 Durand, S., Evangelista, H.B.A., FitzJohn, R., Friendly, M., Furneaux, B., Hannigan,
734 G., Hill, M.O., Lahti, L., Martino, C., McGlenn, D., Ouellette, M.-H., Ribeiro Cunha,
735 E., Smith, T., Stier, A., Ter Braak, C.J.F., Weedon, J., 2025. *vegan: Community
736 ecology package (manual)*.

737 Olivier, P., Frelat, R., Bonsdorff, E., Kortsch, S., Kröncke, I., Möllmann, C., Neumann, H.,
738 Sell, A.F., Nordström, M.C., 2019. Exploring the temporal variability of a food web
739 using long-term biomonitoring data. *Ecography* 42, 2107–2121.
740 <https://doi.org/10.1111/ecog.04461>

741 Patz, J.A., Daszak, P., Tabor, G.M., Aguirre, A.A., Pearl, M., Epstein, J., Wolfe, N.D.,
742 Kilpatrick, A.M., Foufopoulos, J., Molyneux, D., Bradley, D.J., Members of the
743 Working Group on Land Use Change Disease Emergence, 2004. *Unhealthy
744 Landscapes: Policy Recommendations on Land Use Change and Infectious Disease
745 Emergence*. *Environ Health Perspect* 112, 1092–1098.
746 <https://doi.org/10.1289/ehp.6877>

747 Pereira-Silva, J.W., Ríos-Velásquez, C.M., Lima, G.R.D., Marialva Dos Santos, E.F.,
748 Belchior, H.C.M., Luz, S.L.B., Naveca, F.G., Pessoa, F.A.C., 2021. Distribution and
749 diversity of mosquitoes and Oropouche-like virus infection rates in an Amazonian
750 rural settlement. *PLoS ONE* 16, e0246932.
751 <https://doi.org/10.1371/journal.pone.0246932>

- 752 Perrin, A., Glaizot, O., Christe, P., 2022. Worldwide impacts of landscape anthropization on
753 mosquito abundance and diversity: A meta-analysis. *Global Change Biology* 28,
754 6857–6871. <https://doi.org/10.1111/gcb.16406>
- 755 Petruff, T.A., McMillan, J.R., Shepard, J.J., Andreadis, T.G., Armstrong, P.M., 2020.
756 Increased mosquito abundance and species richness in Connecticut, United States
757 2001–2019. *Sci Rep* 10, 19287. <https://doi.org/10.1038/s41598-020-76231-x>
- 758 Poisot, T., Gravel, D., 2014. When is an ecological network complex? Connectance drives
759 degree distribution and emerging network properties. *PeerJ* 2, e251.
760 <https://doi.org/10.7717/peerj.251>
- 761 Ribeiro, S.P., Castro E Silva, A., Dáttilo, W., Reis, A.B., Góes-Neto, A., Alcantara, L.C.J.,
762 Giovanetti, M., Coura-Vital, W., Fernandes, G.W., Azevedo, V.A.C., 2020. Severe
763 airport sanitarian control could slow down the spreading of COVID-19 pandemics in
764 Brazil. *PeerJ* 8, e9446. <https://doi.org/10.7717/peerj.9446>
- 765 Rivera-Hutinel, A., Bustamante, R.O., Marín, V.H., Medel, R., 2012. Effects of sampling
766 completeness on the structure of plant–pollinator networks. *Ecology* 93, 1593–1603.
767 <https://doi.org/10.1890/11-1803.1>
- 768 Rohr, J.R., Barrett, C.B., Civitello, D.J., Craft, M.E., Delius, B., DeLeo, G.A., Hudson, P.J.,
769 Jouanard, N., Nguyen, K.H., Ostfeld, R.S., Remais, J.V., Riveau, G., Sokolow, S.H.,
770 Tilman, D., 2019. Emerging human infectious diseases and the links to global food
771 production. *Nat Sustain* 2, 445–456. <https://doi.org/10.1038/s41893-019-0293-3>
- 772 Santiago-Alarcon, D., Havelka, P., Schaefer, H.M., Segelbacher, G., 2012. Bloodmeal
773 Analysis Reveals Avian Plasmodium Infections and Broad Host Preferences of
774 Culicoides (Diptera: Ceratopogonidae) Vectors. *PLoS ONE* 7, e31098.
775 <https://doi.org/10.1371/journal.pone.0031098>

- 776 Shah, H.A., Huxley, P., Elmes, J., Murray, K.A., 2019. Agricultural land-uses consistently
777 exacerbate infectious disease risks in Southeast Asia. *Nat Commun* 10, 4299.
778 <https://doi.org/10.1038/s41467-019-12333-z>
- 779 Silk, M.J., Croft, D.P., Delahay, R.J., Hodgson, D.J., Boots, M., Weber, N., McDonald, R.A.,
780 2017. Using Social Network Measures in Wildlife Disease Ecology, Epidemiology,
781 and Management. *BioScience* 67, 245–257. <https://doi.org/10.1093/biosci/biw175>
- 782 Smallegange, R.C., Schmied, W.H., Van Roey, K.J., Verhulst, N.O., Spitzen, J., Mukabana,
783 W.R., Takken, W., 2010. Sugar-fermenting yeast as an organic source of carbon
784 dioxide to attract the malaria mosquito *Anopheles gambiae*. *Malar J* 9, 292.
785 <https://doi.org/10.1186/1475-2875-9-292>
- 786 Stone, C., Gross, K., 2018. Evolution of host preference in anthropophilic mosquitoes. *Malar*
787 *J* 17, 257. <https://doi.org/10.1186/s12936-018-2407-1>
- 788 Su, M., Jiang, Z., Hui, C., 2022. How Multiple Interaction Types Affect Disease Spread and
789 Dilution in Ecological Networks. *Front. Ecol. Evol.* 10, 862986.
790 <https://doi.org/10.3389/fevo.2022.862986>
- 791 Takken, W., Verhulst, N.O., 2013. Host Preferences of Blood-Feeding Mosquitoes. *Annu.*
792 *Rev. Entomol.* 58, 433–453. <https://doi.org/10.1146/annurev-ento-120811-153618>
- 793 Toju, H., Tanabe, A.S., Sato, H., 2018. Network hubs in root-associated fungal
794 metacommunities. *Microbiome* 6, 116. <https://doi.org/10.1186/s40168-018-0497-1>
- 795 Valdovinos, F.S., Ramos-Jiliberto, R., Flores, J.D., Espinoza, C., López, G., 2009. Structure
796 and dynamics of pollination networks: the role of alien plants. *Oikos* 118, 1190–1200.
797 <https://doi.org/10.1111/j.1600-0706.2009.17364.x>
- 798 Van Hoesel, W., Marzal, A., Magallanes, S., Santiago-Alarcon, D., Ibáñez-Bernal, S.,
799 Renner, S.C., 2019. Management of ecosystems alters vector dynamics and

- 800 haemosporidian infections. *Sci Rep* 9, 8779. <https://doi.org/10.1038/s41598-019->
801 45068-4
- 802 Villard, M., Metzger, J.P., 2014. REVIEW: Beyond the fragmentation debate: a conceptual
803 model to predict when habitat configuration really matters. *Journal of Applied*
804 *Ecology* 51, 309–318. <https://doi.org/10.1111/1365-2664.12190>
- 805 Visa Shalini, P., Shriram, A.N., Elango, A., Natarajan, R., Vijayakumar, B., Raju, K.H.K.,
806 Dengel, L., Gunasekaran, K., Kumar, A., 2022. Mosquito Diversity in an
807 Experimental Township in Tamil Nadu, India. *Journal of Medical Entomology* 59,
808 1615–1624. <https://doi.org/10.1093/jme/tjac064>
- 809 Vittor, A.Y., Pan, W., Gilman, R.H., Tielsch, J., Glass, G., Shields, T., Sánchez-Lozano, W.,
810 Pinedo, V.V., Salas-Cobos, E., Flores, S., Patz, J.A., 2009. Linking deforestation to
811 malaria in the Amazon: characterization of the breeding habitat of the principal
812 malaria vector, *Anopheles darlingi*. *Am J Trop Med Hyg* 81, 5–12.
- 813 Vizentin Bugoni, J., Maruyama, P.K., Debastiani, V.J., Duarte, L.D.S., Dalsgaard, B.,
814 Sazima, M., 2016. Influences of sampling effort on detected patterns and structuring
815 processes of a Neotropical plant–hummingbird network. *Journal of Animal Ecology*
816 85, 262–272. <https://doi.org/10.1111/1365-2656.12459>
- 817 Wheeler, D., Bhagwat, M., 2007. BLAST QuickStart: Example-Driven Web-Based BLAST
818 Tutorial, in: *Comparative Genomics*. Humana Press, New Jersey, pp. 149–176.
819 <https://doi.org/10.1385/1-59745-514-8:149>
- 820 Young, K.I., Buenemann, M., Vasilakis, N., Perera, D., Hanley, K.A., 2021. Shifts in
821 mosquito diversity and abundance along a gradient from oil palm plantations to
822 conterminous forests in Borneo. *Ecosphere* 12, e03463.
823 <https://doi.org/10.1002/ecs2.3463>