

Reconciling resilience across ecological systems, species, and subdisciplines

Editorial of the cross-journal Special Feature “Reconciling resilience across
ecological systems, species, and subdisciplines”, for consideration for publication in
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Pol Capdevila^{1,2}, Iain Stott³, Imma Oliveras Menor⁴, Daniel B. Stouffer⁵, Rafael L. G.
Raimundo⁶, Hannah White⁷, Matthew Barbour⁸, Roberto Salguero-Gómez^{2,9}

¹School of Biological Sciences, University of Bristol, 24 Tyndall Ave, BS8 1TQ, Bristol, UK

²Zoology Department, Oxford University, Zoology Research and Administration Building, 11a Mansfield
Rd, Oxford OX1 3SZ, UK

³School of Life Sciences, University of Lincoln, Brayford Pool, Lincoln LN6 7TS, UK

⁴Environmental Change Institute, School of Geography and the Environment, University of Oxford,
South Parks Rd OX13QY Oxford, UK.

⁵Centre for Integrative Ecology, School of Biological Sciences, University of Canterbury, Christchurch
8041, New Zealand

⁶Departamento de Engenharia e Meio Ambiente, Programa de Pós-Graduação em Ecologia e
Monitoramento Ambiental (PPGEMA), Centro de Ciências Aplicadas e Educação, Universidade
Federal da Paraíba - Campus IV, Rua da Mangueira s/n, 58297-000, Rio Tinto, PB, Brazil.

⁷School of Life Sciences, Anglia Ruskin University, Cambridge, UK

⁸Department of Evolutionary Biology and Environmental Studies, University of Zurich, 190
Winterthurerstrasse, Zurich 8057, Switzerland

⁹Max Planck Institute for Demographic Research, Rostock Konrad Zuse strasse 1, Rostock 18057,
Germany

Abstract

1. Resilience has emerged as a key concept in ecology and conservation biology to understand and predict ecosystem responses to global change. In its broadest sense, resilience describes the ability of an ecosystem to resist, and recover from, a disturbance. However, the application of such a concept in different sub-disciplines of ecology and in different study systems has resulted in a wide disparity of definitions and ways of quantifying resilience.
2. This Special Feature, which spans the *Journal of Ecology*, *Journal of Animal Ecology* and *Functional Ecology*, provides an overview of how ecologists define, quantify, compare and predict resilience across different study systems.
3. The 29 contributions to this Special Feature show the broad range of approaches used by ecologists to study resilience. Almost half of the contributions (48%) study resilience at the community level, with a 30% of them studying resilience at multiple levels of biological organisation. A large proportion of these articles are observational (42%), experimental (14%) or a combination of both (17%), whilst a 17% utilise theoretical or computational approaches. Whilst 38%, 21% and 14% of the studies were based solely on plants, animals or microorganisms respectively, 17% of them incorporated these multiple trophic levels.
4. *Synthesis*. A unified ecological understanding of resilience across systems and taxa requires a trans-disciplinary consensus on what resilience actually is and how to best measure it. Here, we provide an overview of how ecologists define, quantify, compare, and predict resilience across different ecological systems and subdisciplines, with reference to the diverse approaches used by contributions to this Special Feature. We identify four key recommendations to

harmonise future efforts in resilience research: (1) define resilience using existing theoretical frameworks; (2) use common and comparable metrics to measure resilience; (3) clearly contextualise and define the pre- and post-disturbance state of the ecological system; (4) consider explicitly the disturbance type and regime impacting the system.

Keywords: Conservation biology, Community, Disturbance, Global change, Population, Recovery, Regime shift, Resistance, Stability.

Resumen (Español)

1. La resiliencia ha surgido como un concepto clave en ecología y biología de la conservación para comprender y predecir las respuestas de los ecosistemas al cambio global. En su sentido más amplio, la resiliencia describe la capacidad de un ecosistema para resistir y recuperarse de una perturbación. Sin embargo, la aplicación de este concepto en diferentes subdisciplinas de la ecología y en diferentes sistemas de estudio ha resultado en una gran variedad de definiciones y formas de cuantificar la resiliencia.
2. Este edición especial, que comprende las revistas *Journal of Ecology*, *Journal of Animal Ecology* y *Functional Ecology*, proporciona una descripción general de cómo los ecólogos definen, cuantifican, comparan y predicen la resiliencia en diferentes sistemas de estudio.
3. Las 29 contribuciones a esta edición especial muestran la amplia gama de enfoques utilizados por los ecologistas para estudiar la resiliencia. Casi la mitad de las contribuciones (48%) estudian la resiliencia a nivel de comunidades ecológicas, y un 30% de ellos estudia la resiliencia en múltiples niveles de organización biológica. Una gran proporción de estos artículos son

observacionales (42%), experimentales (14%) o una combinación de ambos (17%), mientras que un 17% utiliza enfoques teóricos o computacionales. Mientras que el 38%, 21% y 14% de los estudios se basaron únicamente en plantas, animales o microorganismos, respectivamente, el 17% de ellos incorporaron estos múltiples niveles tróficos.

4. *Síntesis*. Una comprensión ecológica coherente de la resiliencia entre sistemas y taxones requiere un consenso transdisciplinario sobre qué es realmente la resiliencia y cómo medirla. En esta editorial proporcionamos una descripción general de cómo los ecólogos definen, cuantifican, comparan y predicen la resiliencia en diferentes sistemas ecológicos y subdisciplinas, con referencia a los diversos enfoques utilizados por las contribuciones a esta edición especial. Identificamos cuatro recomendaciones clave para armonizar los esfuerzos futuros en la investigación de la resiliencia: (1) definir la resiliencia utilizando los marcos teóricos existentes; (2) utilizar métricas comunes y comparables para medir la resiliencia; (3) contextualizar y definir claramente el estado anterior y posterior a la perturbación del sistema ecológico; (4) considere explícitamente el tipo de perturbación y el régimen que impacta el sistema.

95 **Introduction**

96 As the impacts of global change continue to unfold worldwide (Díaz et al., 2019; IPCC,
97 2021; Maxwell et al., 2016), understanding the ability of ecological systems to respond
98 to global threats has become a pressing societal need (CBD, 2010; UNISDR, 2015).
99 In recent decades, the anthropogenic stressors impacting ecological systems have
100 escalated at unprecedented rates in both number and severity (Barnosky et al., 2012;
101 Díaz et al., 2019; Newbold et al., 2015). Despite efforts to prevent global change
102 impacts, species extinctions have increased 100- to 1,000-fold (Barnosky et al., 2011;
103 Ceballos et al., 2015), with approximately 1 million species predicted to become extinct
104 over the coming decades (Scholes et al., 2018). The on-going loss of species is
105 altering the structure and functioning of ecosystems worldwide (Pecl et al., 2017). As
106 a consequence, preserving resilience, defined as the ability of ecological systems to
107 resist and recovery from disturbances (Hodgson et al., 2015), has become a key
108 conservation priority. For instance, several international environmental policies, such
109 as the Aichi Biodiversity Targets (CBD, 2010), the Sustainable Development Goals
110 (United Nations General Assembly, 2015) or the Sendai Framework (UNISDR, 2015)
111 explicitly include preserving resilience as a target.

112 Despite its importance in ecology and conservation, the popularity of resilience
113 is rivalled only by the disparity in its interpretations, definitions, and applications across
114 different ecological subdisciplines (Hodgson et al., 2015; Ingrisch & Bahn, 2018).
115 These discrepancies between approaches mean that ecologists have used a variety
116 of different ‘indicators’ or ‘metrics’ of resilience (Angeler & Allen, 2016). As such,
117 *Journal of Animal Ecology*, *Journal of Ecology*, and *Functional Ecology* present this
118 joint Special Feature to provide an updated overview of the different ways ecologists
119 define, quantify, compare, and predict resilience across different ecological systems,

species, and subdisciplines. In this Editorial, we discuss current perspectives on resilience, and both theoretical and empirical approaches to studying it, across the 29 manuscripts published in this Special Feature. We first provide a brief conceptualisation of resilience and ways of quantifying it, linking to the approaches used in this Special Feature. Second, we identify the key opportunities and challenges to advance our understanding of resilience in ecological systems, and describe how the Special Feature papers contribute in these new directions. Finally, we propose several steps to move the field further towards an integrated understanding of resilience across ecological scales.

The concept of resilience across subdisciplines

Resilience (Lt. '*resilire*', to leap or spring back, OED, 1989) is a widely used concept in ecology. Over the course of history, different scientific disciplines have adopted the term resilience to describe different processes. For example, the physician James Carson (1820) used the term to describe the ability of lungs to expand and contract. In psychology, resilience was first coined to describe the capacity of children to endure difficult emotional experiences (Rutter, 1979). In engineering, resilience was first used to describe the stress (in terms of load bearing weight) that timber could sustain before breaking (Tredgold, 1818). Despite the disparity of the use of resilience in different fields, the commonality among them is that *resilience describes the capacity of a system to deal with change*.

Given the tight link between resilience and change, it is not surprising that resilience has become fundamental to ecological research. Early understanding of resilience was predicated on ecosystem stability and persistence despite disturbances (e.g. MacArthur, 1955), where disturbances represent a/biotic factors displacing the

system away from its stationary equilibrium (Holling, 1973). Ecological systems were understood to persist in determined states, representing stable and equilibrium conditions to which the system is assumed to return back to following any disturbance (Lewontin, 1969; May, 1977). Though not explicitly linked to the term resilience *per se*, these early works on stability introduced crucial concepts, such as alternative stable states or basins of attraction, which set the foundations of ecological resilience theory.

The first formalisation of the term resilience in ecology - that we are aware of - was made by Holling (1973), although he already used the term in some of his earlier studies (e.g. Holling & Goldberg, 1971; Holling & Orians, 1971). Holling suggested that the stability of an ecosystem is a different property than its resilience and argued that ecological systems can exist in multiple, alternative stable states (Holling, 1996; Figure 1). Each state is defined by different stable system structures, compositions, and processes, maintained by forces or feedbacks that represent their basin of attraction (Figure 1; Folke et al., 2004; Holling, 1973, 1996). When a disturbance displaces the system from its current state, feedback processes draw it back to that initial state (Figure 1). If the system is disturbed beyond the basin of attraction and past a tipping point, however, it may undergo a regime shift and transition to an alternative stable state (Dakos et al., 2014; Scheffer et al., 2009). As such, Holling defined resilience as *“the size of a stability domain or the amount of disturbance a system could take before it shifted into alternative configuration”* (*sensu* Holling, 1973).

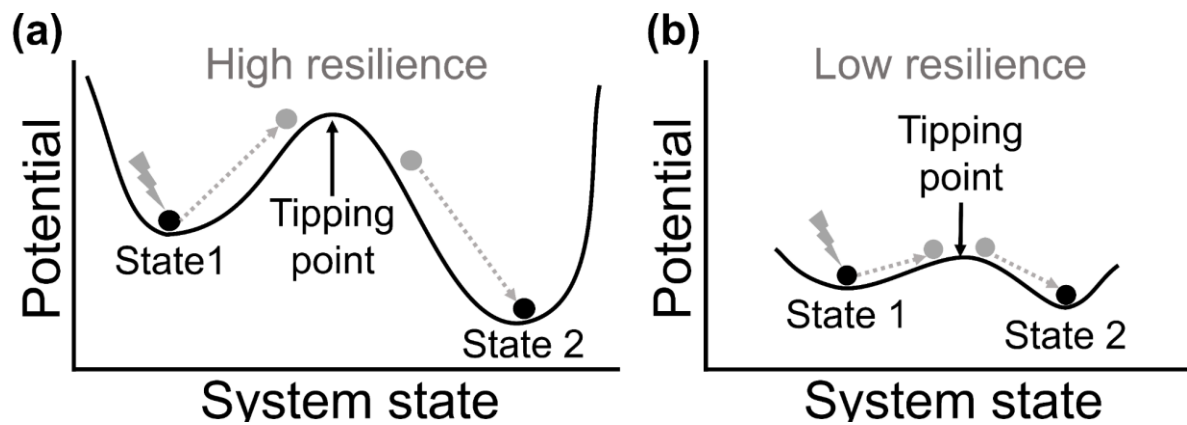


Figure 1. Classical analogy of the rolling ball to represent the resilience of an ecological system according to Holling (1973). Each black ball represents the stable state of the system, whereas the grey balls illustrate the potential trajectory of the system. The basins represent the potential system states. Lightning symbols represent a disturbance to the system. To shift from a stable state to another one, the system must be perturbed sufficiently by a disturbance (lightning) for the ball to surpass a tipping point and roll over from one basin to another (dashed arrow). The width and depth of the basin are related to resilience: a system with a deep and wide pit (a) will be more resilient than a system with a flat and narrow pit (b), given that a more extreme disturbance will generally be needed to cause a regime shift.

Due to the rise of discrepancies around the concept of resilience (e.g. Pimm, 1984), Holling distinguished two main approaches to quantify resilience: *engineering resilience* and *ecological resilience* (Holling, 1996). Engineering resilience defines resilience as the process of *recovery of the system* following a disturbance (Pimm, 1984). This view considers resilience to be a component of system stability (Donohue et al., 2013; Pimm, 1984), and often assumes that a system can only have a single stability regime (Holling, 1996). On the other hand, ecological resilience assumes that a system may have multiple alternative states and defines resilience as *resistance* to change, *i.e.* the magnitude of disturbance that a system can absorb before shifting from one state to another (Holling, 1996). These definitions of ecological vs. engineering resilience have diverged substantially in the ecological literature (Brand & Jax, 2007). More recent views consider that resilience encompasses multiple

components describing both resistance *and* recovery as described above (Hodgson et al., 2015; Ingrisch & Bahn, 2018).

Measuring resilience across ecological systems

Whilst theory helps to conceptualise phenomena such as resilience, empiricism requires specifics. This need is precisely what the title of this Special Feature alludes to: “Reconciling resilience across ecological systems, species, and subdisciplines”. Ecological systems operate at different levels of biological organisation (individuals, populations, communities), which correspond to various ecological subdisciplines. All ecological systems have a structure formed of interacting system components (e.g. age structure of a population, Caswell, 2001; functional composition of species in a food web, Ings et al., 2009). All systems function in terms of how components interact, such as interaction between species (e.g. predation pressure, Donohue et al., 2017), progression through life history stages (e.g. maturation rate, Stearns, 1992), or interaction between individuals (e.g. breeding effort, Ricklefs, 1977). All systems have measurable system outputs, commonly including size (e.g. population size), growth (e.g. rate of community biomass change), diversity (e.g. species richness), or composition (e.g. sex ratio in a population). System outputs are also often measured using functional traits (Carmona et al., 2016; Violle et al., 2007), at various output levels (individual, population, community). To quantify resilience (i.e. the capacity of a system to deal with change), one needs to (1) consider how structure and function are defined, as well as in which ways these are affected by disturbance (i.e. “change” imposed on the system), (2) identify relevant measures of system outputs (i.e. the best understanding of system “capacity”), and (3) develop metrics to quantify those outputs at the appropriate level of organisation, which can vary both between and within study

species and systems (i.e. best ways to measure capacity of the system to respond to change).

In population ecology, resilience has been studied for decades (Harrison, 1979; Neubert & Caswell, 1997), though not necessarily with explicit recourse to resilience theory (e.g. Stott et al., 2011). Population models are typically formulated using the (st)age structure of the population, with the life cycle of the species defined by average vital rates (e.g. survival, development, reproduction) clustered into (st)ages. Here, models assume that populations display stable states defined by population structure - i.e. the relative number of individuals in each life cycle (st)age, with commensurate stable numerical growth or size (Caswell, 2001). Simple (density-independent and non-stochastic) conceptualisations assume that populations converge to a stable state defined by the relative proportions of life cycle (st)ages and maintained by the vital rates of the population (Caswell, 2001). However, populations are frequently subject to disturbances that displace them from their stable structure, thus changing the relative proportions of individuals with high *versus* low survival and/or fecundity (Caswell, 2001; Stott et al., 2011). Resistance and return rate of a structured population can thus be measured relative to its pre-disturbance population size, growth, and/or structure (Caswell, 2001; Stott et al., 2011). Despite the clear links between these concepts and resilience theory, the connections were not formalised until recently (Capdevila et al., 2020). Similar conceptualisations could reasonably be applied to equilibrium states in density-dependent population dynamics, and expected growth in stochastic population dynamics.

Resilience in communities has received a considerable amount of attention in ecological research. Community “structure” is often understood as the network of interactions of species (*sensu* Caswell, 1976), but could also be understood in terms

of “composition” using taxonomic, phylogenetic or functional groupings of species (Carmona et al., 2016; Pérez-Valera et al., 2018). Community structure and composition may be measured in a multitude of ways, such as measures of species richness, numerical abundance, biomass, or phylogenetic diversity (Ings et al., 2009; Tylianakis et al., 2008). Early theory assumed communities to be in a stable state in terms of absolute or relative species abundance, with species interactions being the ‘processes’ underlying community function (May, 1977). The ‘state’ from which a community departs or to which it may return may be linked to those measures mentioned earlier in this paragraph or others (e.g. Cole et al., 2014; Yang et al., 2019). Though earlier works often considered limited interaction types (the trophic interactions of food webs being probably the most studied), contemporary research is increasingly concerned with different interaction types and their relative strengths (Li et al., 2021). A great deal of debate still abounds about the effects of the complexity of the community network on its stability and resilience: while early theoretical works showed complexity to be destabilising (Magurran, 2013), more recent theory has shown that complexity can be stabilising (Mougi & Kondoh, 2012; Qian & Akçay, 2020).

While most resilience approaches have focused on quantifying changes in the structure and composition of ecological systems (Hughes et al., 2003; Lloret et al., 2011), an emerging area of research is to quantify the resilience of ecosystem functions (Oliver et al., 2015). Focusing on community structure and/or composition to examine resilience risks rendering an incomplete picture of the extent of the impacts that disturbances might have on the functionality of ecological systems (Gladstone-Gallagher et al., 2019; Matos et al., 2020; Oliver et al., 2015). Species contributions to ecosystems functions are tightly linked to their functional traits —morphological,

physiological, phenological, or behavioural features, measurable at the individual level, that have an impact on species fitness (*sensu* Violle et al., 2007). Different species might share similar combinations of traits and so provide similar or equivalent ecosystem functionality, *i.e.* be functionally redundant (Carmona et al., 2016; de Bello et al., 2010). Consequently, similar ecosystem functioning might be achieved by different communities (Gallagher et al., 2013), illustrating the possible independence between structure and composition of ecosystems, and their functionality. This need to distinguish between composition and functionality gave rise to the concept of *functional resilience* (Oliver et al., 2015). Functional resilience, “*the degree to which an ecosystem function can resist or recover rapidly from environmental perturbations*” (*sensu* Oliver et al., 2015), incorporates a more recent view of resilience.

Contributions to this Special Feature

This Special Feature comprises 29 pieces covering a broad range of topics related to resilience of ecological systems. To better contextualise them, we classified these contributions according to the ecological subdiscipline, the approach, and the system through which the authors study resilience (Figure 2). Perhaps as a legacy of early works (Holling, 1973; Pimm, 1984), almost half of the contributions in this Special Feature examine resilience at the level of communities (Figure 2a). Yet, a number of contributions cut across different levels of organisation, with a large proportion of the works studying resilience levels at the interfaces of community, population and functional perspectives (Figure 2a). The patterns that emerge from this set of publications suggest that ecological research is making progress in breaking previous legacies of focusing only on one level of biological organisation, though more interdisciplinary opportunities lie ahead, as we discuss in the final section of this editorial.

Early developments of resilience research were mostly based on theoretical works (e.g. Harrison, 1979; Holling, 1973). These theoretical approaches have sometimes been challenging to apply to “real world” systems, hampering the use of resilience in applied disciplines (Hodgson et al., 2015; Ingrisch & Bahn, 2018). In this Special Feature, most of the contributions are observational studies, with a large proportion combining experimental and observational approaches (Figure 2b). Such a combination of approaches is important to provide a better understanding of the resilience of ecological systems. For example, experiments might help to unravel processes promoting resilience, which might be more difficult to observe in the field (Hoover et al., 2021; Jones et al., 2020; Lipoma et al., 2021). Combining theoretical studies with experiments and/or observational studies can also help to test the validity of resilience concepts (Li et al., 2021; Medeiros et al., 2021).

Because natural systems are complex, it is challenging to study all their components. In this Special Feature, most of the studies focus on plants or algae, with a relatively large proportion studying resilience by including multiple taxonomic groups, though these are mostly theoretical studies (Figure 2c). While focusing on a single system can simplify experimental and observational studies (e.g. only primary producers in a community), where feasible, incorporating multiple components of the system (e.g. further trophic levels) will render a better understanding of it as a whole.

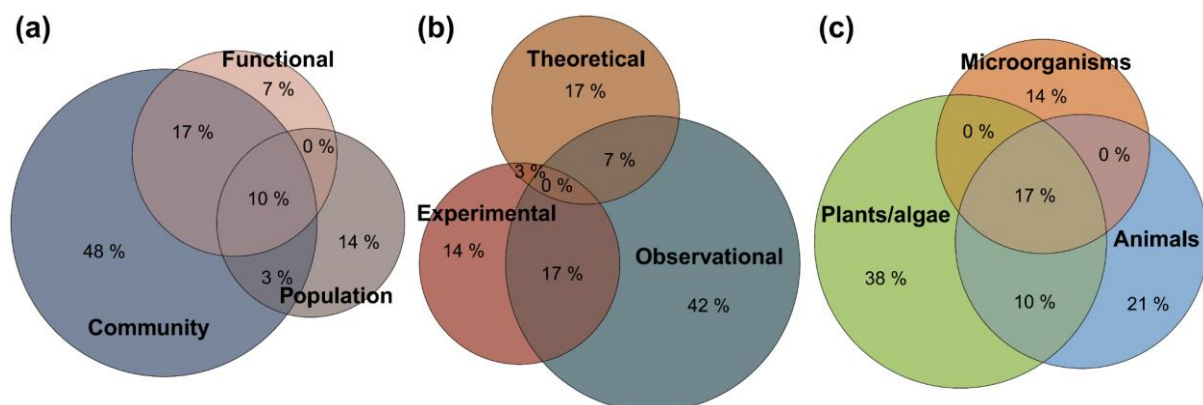


Figure 2. Venn diagrams showing the domain of operation of the 29 manuscripts included in the joint Special Feature “Reconciling resilience across ecological systems, species, and subdisciplines”. The different panels show the proportion of contributions according to: **(a)** the ecological subdisciplines, including population ecology, community ecology, and functional ecology; **(b)** whether the approach used to study resilience was experimental, theoretical, and/or observational approaches; **(c)** the species or taxa studied, classified as plants/algae, animals, and/or microorganisms. Conceptual contributions are assigned to the three ecological subdisciplines in panel **a** and the three taxonomic groups in panel **c** due to their cross-applicability.

Opportunities and challenges in the Special Feature

Conceptualising and operationalising resilience

A unified ecological understanding of resilience across systems and species requires a trans-disciplinary consensus on what resilience actually is and how to best measure it. As such, one of the most pressing challenges ahead is to bring consensus across traditionally disparate fields. In this Special Feature, two key contributions, Delettre (2021) and Van Meerbeek et al. (2021), provide complementary perspectives on current definitions and quantifications of views of resilience, given historical divergences in its study. Both contributions recognise the importance of semantics in furthering the study of resilience, and make key suggestions regarding ways to overcome conflicting definitions. Delettre (2021) stresses that the different concepts of resilience used in the literature do not represent degrees of resilience but rather types of behaviour of a system. Hence, it is important to choose the appropriate definition of resilience depending on the research questions or management goals, the processes by which persistence is achieved, and the types of disturbance and spatio-temporal scales considered. Semantics aside, understanding commonalities across systems in terms of how they are structured, how they function, and how they can be measured, may be a good start towards reconciling how we ecologists study resilience.

One casualty of diverse and divergent parlances is an inability to compare results between studies asking similar questions. Synthesis and meta-analysis in ecology have proven extremely powerful tools (Koricheva & Gurevitch, 2014); but the current state of the art in resilience research is one that makes finding generality challenging. As pointed out by Van Meerbeek et al. (2021), it would be advantageous to standardise metrics depicting resilience/stability components to facilitate comparisons across studies. Over the last decades, ecological research has made significant methodological advances, with an increasing number of statistical techniques (e.g. time-series analyses, Dennis et al., 2006; network analyses, Blüthgen, 2010; spatial analyses, Dale & Fortin, 2014), and sharing of methods and tools will be advantageous to everyone. Using common “currencies” of the components of resilience/stability (e.g. Capdevila et al., 2020; Clark et al., 2021; Ingrisch & Bahn, 2018) -or at least determining “conversion factors” across different metrics - will make comparisons among studies possible, opening up the possibility of much-needed global assessments of resilience.

Understand resilience under different disturbance regimes

Resilience research is inherently linked to the properties of the disturbances altering ecological systems (Bender et al., 1984). Historically, resilience research has predominantly focused on sudden events or pulse disturbances (Holling, 1973; Pimm, 1984). However, disturbances can occur at different intensities (low to high) and frequencies (pulse to press) (Jentsch & White, 2019). Hence, accounting for different disturbance regimes is crucial to understand the resilience of ecological systems to global change, particularly so because a change in the natural disturbance regime can have profound impacts on the systems’ resilience. To this end, two contributions to this Special Feature demonstrate that changes in sea temperature, acting as a chronic

stressor, can have major impacts on the functioning (Tsimara et al., 2021) and stability (Miner et al., 2021) of marine communities. Tsimara et al. (2021) combine data on Mediterranean fisheries landings over 31 years (1985-2015) and species traits to infer the resilience dynamics and build stability landscapes. On the other hand, Miner et al. (2021) utilise a decade-long data set of rocky intertidal communities from the whole of the U.S. West coast, to quantify the temporal and spatial community changes as a measure of stability. Likewise, Serra et al. (2021) explore the influence of vegetation clearing frequency and forest age on the recovery. The authors quantify resilience as a measure of the recovery of the number and diversity of soil macrofauna in the Brazilian Amazon.

The effects of multiple disturbances often compound one another in ecological systems (Côté et al., 2016; Orr et al., 2020). A number of contributions to this Special Feature explore the interactive effects of multiple disturbances on systems. Lipoma et al. (2021) explore the interactive effects of land-use and weather variability on the rate of change of vegetation towards the primary forest, which they consider to be the reference state of the system. Their findings suggest that long-term land use might induce long-term changes in the ecological system hampering their engineering resilience in a short period of time (five years after disturbance cessation). Nowicki et al. (2021) highlight how trophic cascades, through the loss of predators, can exacerbate the impacts of extreme climatic events on the temporal changes in seagrasses and macroalgae cover in Western Australia, using initial cover as reference state. Similarly, Nelson et al. (2021) report the interactive effects of warming and drought on both resistance and recovery of invertebrate community abundance in food webs with different energy channel configurations. Their results suggest that communities inhabiting streams with large amounts of organic matter and more

complex substrates are more resilient to the loss of surface water than communities inhabiting streams with simpler, more homogeneous substrates.

Accounting for the temporal and spatial scale at which disturbances occur is also crucial. A large proportion of the ecological literature has focused on studying the immediate response of ecological systems to disturbances (e.g. Cole et al., 2014; de Vries et al., 2012; DeSoto et al., 2020). Yet, the effects of such disturbance on the system might take several years or even decades to manifest (Hughes et al., 2013; Johnstone et al., 2016). In this Special Feature, a number of contributions show the importance of the “legacy effects” of disturbances into the present resilience of ecological systems (Johnstone et al., 2016). In this sense, Leizeaga et al. (2020), report a low sensitivity to droughts in bacterial growth, fungal growth and respiration that have been historically affected by high drought frequency in a gradient of precipitation in Texas, USA. Hoover et al. (2021), demonstrate that the long-term effects of seasonal droughts on soil moisture can impact the resistance and recovery of plant biomass and phenology in Colorado, USA.

Webster et al. (2021) show how following extreme rainfall events the resistance to changes in biomass and leaf density of seagrass populations, and the recovery to pre-disturbance historical values, depends on the salinity levels to which these have been exposed previously. Leverkus et al. (2020) use a meta-analysis approach to highlight how local environmental factors also play a key role in the resilience of trees to logging. Also, Ovenden et al. (2021) report a high sensitivity of different metrics of forest resilience to the period of time considered as baseline, which calls into caution the need for a clear definition of the stable state of the system under examination. Finally, Steel et al. (2021) show that topography and vegetative structure influence of

410 on the resistance and recovery of forest vegetation cover and heterogeneity in
411 California's Sierra Nevada mountain.

412 *Integrating multiple levels of biological organisation*

413 Ecological systems are often studied at different levels of biological organisation
414 (individuals, populations, communities). However, by examining resilience in the
415 context of processes happening at specific levels of organisation, we may miss
416 important drivers of a system's resilience emerging from bottom-up or top-down
417 processes in constituent sub-systems or overarching super-systems. In this Special
418 Feature, a number of contributions provide key examples of how to integrate data,
419 framework, and methods to examine resilience in a holistic manner. Lisovski et
420 al.(2020) show that specific traits can impact resilience at a population level: migration
421 behaviours in two shorebird species differentially affect individual survivorship, which
422 has implications for population resilience. Populations lie at the intersection between
423 processes that directly shape individual and community performance (Griffith et al.,
424 2016). In this context, Paniw et al. (2021) show a high degree of complementarity
425 between demographic and functional traits in facilitating community composition and
426 cover resilience to droughts. For example, vital rates are more important in explaining
427 total and individual species resilience, while functional traits matter more to explain
428 compositional resilience.

429 Carnicer et al. (2021) combine ecophysiological and demographic metrics to
430 determine the resilience of sessile oaks (*Quercus petraea*) to droughts and
431 heatwaves. A great deal of population variation was found regarding individual
432 secondary growth, recruitment, and thermal exposure of saplings to heatwaves,
433 mostly driven by microhabitat conditions. The authors use 20 different resilience,

resistance and recovery indices comparing secondary growth before and after disturbance. Muñoz et al. (2021) combine demographic and community data to show that the resilience of tropical forests is driven by autogenic regulation. The authors used long-term community data from old-growth and secondary forests in southern Mexico to analyse three key state variables (basal area, tree density, species richness), their annual rates of change, and their underlying demographic processes (recruitment, growth, mortality). They find a negative relationship between state variables, their rates of change and their underlying demographic processes, supporting that forest dynamics is driven by autogenic factors.

Unravelling the relationship between the multiple components of resilience

Because of the multifaceted nature of resilience, a key question ahead is whether and how its components are related to each other. In this Special Feature, Medeiros et al. (2021) reveal that recovery and resistance are negatively correlated with one another using both experimental microbial systems and theoretical models, suggesting that resistance could be inferred from recovery and *vice versa*. Likewise, Jones et al. (2020) show that both the resistance and recovery of plant communities to pulse disturbances (*i.e.* sudden events) are similarly affected by flooding stress gradients in salt marshes in Louisiana, USA. Moreover, it is key to understanding whether the linkages between the different components of resilience hold when the systems are exposed to disturbances (Donohue et al., 2013). Eagle et al. (2021) demonstrate that flood events can alter the correlations between five different metrics of stability on freshwater macroinvertebrate communities. The authors use a 18-year time series (2000-2017) of macroinvertebrate community dynamics from a southeast Alaskan river, illustrating how stability can be examined in natural ecosystems time series data.

What makes a system resilient?

A key challenge in ecology is to predict the resilience of ecological systems to future, and potentially novel, disturbances and environmental conditions (Sutherland et al., 2013). Global threats, such as global warming (IPCC, 2021) or habitat loss (Newbold et al., 2015), are likely to continue to impact ecosystems worldwide even in the most optimistic conservation policy scenarios (Leclère et al., 2020). However, predicting resilience is not an easy task, not only because of the abovementioned discrepancies in the field, but also because it is an emergent property of complex systems (Scheffer et al., 2018). Hence, we need to develop frameworks that can help us to anticipate the potential consequences of the current ongoing global change into the future resilience of ecological systems.

Trait-based approaches could provide a solution to this challenge. Indeed, these approaches are becoming more accessible to ecologists, with standardised protocols for data collection (Moretti et al., 2017) and global databases already at hand (e.g. TRY, Kattge et al., 2020; Amniote, Myhrvold et al., 2015). In this Special Feature, Bonhomme et al. (2020) report that drought applies selection pressures on invertebrate species living within water pools in bromeliad plants, according to feeding traits and ability to tolerate drought stress. They show that resilience, measured using both functional and taxonomic diversity, is more dependent on these traits, and particularly stress-tolerance of resting stages such as eggs and cysts, than on meta-community dynamics of post-disturbance immigration. Su et al. (2020) show that trait-based early warning signals can be used to anticipate both the collapse and the recovery of a lake ecosystem in the Yangtze floodplain to multiple disturbances (warming, eutrophication, and biotic interactions). Studies such as these may be pivotal in informing management, and De Battisti (2021) proposes a conceptual framework for predicting functional resilience of communities. The author illustrates

484 how different suites of plant traits can help predict the resistance and recovery of salt
485 marshes and sand dunes to pulse, chronic, and rapid onset disturbances. De Battisti
486 argues that, by linking plant functional traits to the resilience of coastal ecosystem
487 properties, we can provide actionable plans for resource managers.

488 Some network structures will be more disposed to high resilience than others,
489 and specific “keystone” species, species groups, interactions, cascades or feedback
490 loops may indicate greater capacity for a system to withstand environmental
491 disturbances or change. Maia et al. (2021) use adaptive population-dynamics models
492 to indicate that herbivory networks and their high degree of specialisation show
493 robustness against extinction cascades. Pollination networks, on the other hand, show
494 high generalisation which appears to make them more vulnerable to species loss in
495 the short term. However, their structure confers an adaptive capacity that could be
496 leveraged in efforts designed to restore or maintain key ecosystem functions like
497 pollination. Likewise, Thakur et al. (2021) show that heat shocks applied to
498 rhizosphere microcosms decreased prey biomass to a far greater extent than predator
499 biomass, with prey biomass relatively low through the recovery period. These results
500 highlight how the same disturbance can promote imbalance in the structure of food
501 webs due to differences in the resilience of the components of a system. Li et al. (2021)
502 demonstrate that energetic constraints at the trophic group and food web level
503 enhance resilience by dampening the strength of destabilizing positive feedback loops.
504 Jia et al. (2020) reveal that the presence of arbuscular mycorrhizal fungi (AMF) in
505 grassland ecosystems promotes resistance and improves resilience to drought. AMF
506 aided recovery of the community following drought, and promoted resistance to
507 drought as measured using plant productivity and nitrogen cycling, particularly
508 ameliorating compounding adverse effects of N deposition. Finally, Mungi et al. (2021)

demonstrate that the role of protected areas in providing resistance to species invasions, measured indirectly as the lack of invasive species, is context dependent. The authors use data on plant communities (species richness and abundance) from five tropical forest types inside and outside protected areas, also accounting for other covariates such as climate, forest type, anthropogenic disturbance and native plant richness.

Conclusions

Despite decades of research, important knowledge gaps remain in our understanding regarding the resilience of ecological systems. The contributions to this joint Special Feature address some of these gaps, using a mix of theoretical and empirical means, using natural and experimental case studies, across ecological systems within and across scales of biological organisation. They also naturally open up new and exciting research avenues. For the field of ecological resilience to move forward, we identify four recommendations to harmonise future research efforts.

(1) *Define resilience using existing frameworks.* Existing frameworks currently provide both clear definitions and ways to quantify the resilience and the stability of ecological systems (Capdevila et al., 2020; Donohue et al., 2013; Hodgson et al., 2015; Ingrisch & Bahn, 2018; Oliver et al., 2015). Future studies would benefit from making it clear where their resilience approach sits within the existing resilience frameworks, distinguishing whether they are studying resilience, stability, or any of their sub-components. While distinctions such as ecological vs. engineering resilience have been helpful in the past, contemporary frameworks might provide a more holistic approach to integrate the different components of resilience (Capdevila et al., 2020; Hodgson et al., 2015; Ingrisch & Bahn, 2018). Thus, identifying the variables of interest

and how they are measured, within such frameworks, will help cohesion and comparison across studies.

(2) *Use common metrics to measure resilience.* Studies should aim to measure resilience using standardised metrics that are applicable both in theoretical and empirical studies, and that are comparable among different systems. For example, measuring the relative change in abundance before and after a disturbance could represent a measure of resistance in both communities and populations. In this sense, Ingrisch & Bahn (2018), provide an extensive review on how to standardise measures of resilience across systems. Using a unified approach will facilitate comparisons among different systems and scales of biological organisation (Clark et al., 2021; Ingrisch & Bahn, 2018), as well as linking theoretical and observational studies. Beyond that, common metrics will help to find global patterns of resilience across different systems (e.g. Capdevila et al., 2021), as well as contributing to improve our mechanistic understanding of how ecological systems achieve resilience.

(3) *Define the pre- and post-disturbance state.* Independent of the scale and level of organisation at which resilience is measured, all systems have a given structure and composition with measurable outcomes (e.g. size, diversity). It is then crucial to define such a reference state from which resilience and/or its components will be measured for better contextualisation. For example, if one wants to measure resistance as the ability of the system to remain unchanged after a disturbance, it is crucial to have a reference state of the system before the disturbance. Defining a reference state can be achieved either by characterising the system before the disturbance or by using undisturbed control treatments (Ingrisch & Bahn, 2018).

(4) *Define the disturbance type and regime.* The resilience of a system is sensitive to the kind of disturbance (Bender et al., 1984; Johnstone et al., 2016). It is therefore important to clearly define the nature of the disturbance affecting that determined system. That is, distinguishing whether these are pulse (e.g. storm, fire), or press disturbances (e.g. global warming, ocean acidification). This distinction is important to also define the trajectory of the system towards its “recovered state”. For instance, a chronic disturbance might cause a permanent system change, where a return to stability can only be achieved through adaptation (Hodgson et al., 2015). Furthermore, frequency and intensity of disturbance events will have a strong impact on system recovery, dependent on (non)linearity of system resistance to disturbance intensity, and the recovery time required following a disturbance event as compared to disturbance frequency. Understanding the impact of different disturbance regimes on resilience is therefore particularly important, given existing and expected increases in intensity and frequency of large disturbance events due to climate change (IPCC, 2021).

Resilience is a common component of how we understand the response of the natural world to global threats and change. Moreover, conserving resilience in nature is an explicit goal of global conservation efforts (CBD, 2010; UNISDR, 2015; United Nations General Assembly, 2015). Bringing consensus to how resilience is conceptualised will render a better understanding of resilience across diverse ecological systems by framing it in terms of consistent components. This consensus of course requires clarity in how these components are measured. In doing so, barriers - which in our views are artificial - between ecological subdisciplines, and indeed between different schools of thought within resilience research, may begin to dissolve. This will bring commensurate benefits to ecology as we begin to understand the ripple

effects of resilience up and down systems at different levels of biological organisation. This knowledge will ultimately provide crucial guidance to develop and apply effective management actions, informing where to allocate the inherently limited resources for nature conservation.

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Authors contributions

P.C., I.S., and R.S.-G. set the foundations to the special feature. P.C. coordinated its development and integration. I.O.M., D.B.S., R.L.G.R., H.W. and M.B. summarised most of the contributions to the Special Features and provided feedback to initial and final versions of the manuscript. P.C., I.S. and R.S-G. wrote the first version of the editorial and integrated feedback from co-authors and reviewers. All authors contributed to the article and gave final approval for publication.

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606 **Data availability statement**

607 No new data were used in this manuscript

608

609 **Conflict of interest statemen**

610 Imma Oliveras Menor and Iain Stott are Associate Editors of Journal of Ecology, but
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612 Barbour, Rafael Raimundo, Daniel Stouffer, and Hannah White are Associate Editors
613 and Roberto Salguero-Gómez is Commissioning Editor of Journal of Animal Ecology,
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616 **References**

- 617 Angeler, D. G., & Allen, C. R. (2016). Quantifying resilience. *Journal of Applied*
618 *Ecology*, 53(3), 617–624.
- 619 Barnosky, A. D., Hadly, E. A., Bascompte, J., Berlow, E. L., Brown, J. H., Fortelius,
620 M., Getz, W. M., Harte, J., Hastings, A., Marquet, P. A., Martinez, N. D., Mooers,
621 A., Roopnarine, P., Vermeij, G., Williams, J. W., Gillespie, R., Kitzes, J.,
622 Marshall, C., Matzke, N., ... Smith, A. B. (2012). Approaching a state shift in
623 Earth's biosphere. *Nature*, 486(7401), 52–58.
- 624 Barnosky, A. D., Matzke, N., Tomiya, S., Wogan, G. O., Swartz, B., Quental, T. B.,
625 Marshall, C., McGuire, J. L., Lindsey, E. L., & Maguire, K. C. (2011). Has the
626 Earth's sixth mass extinction already arrived? *Nature*, 471(7336), 51–57.

627 Bender, E. A., Case, T. J., & Gilpin, M. E. (1984). Perturbation experiments in
628 community ecology: Theory and practice. *Ecology*, 65(1), 1–13.

629 Blüthgen, N. (2010). Why network analysis is often disconnected from community
630 ecology: A critique and an ecologist's guide. *Basic and Applied Ecology*, 11(3),
631 185–195.

632 Bonhomme, C., Céréghino, R., Carrias, J.-F., Compin, A., Corbara, B., Jassey, V. E.
633 J., Leflaive, J., Farjalla, V. F., Marino, N. A. C., Rota, T., Srivastava, D. S., &
634 Leroy, C. (2020). In situ resistance, not immigration, supports invertebrate
635 community resilience to drought intensification in a Neotropical ecosystem.
636 *Journal of Animal Ecology*.

637 Brand, F. S., & Jax, K. (2007). Focusing the meaning (s) of resilience: Resilience as a
638 descriptive concept and a boundary object. *Ecology and Society*, 12(1).

639 Capdevila, P., Stott, I., Beger, M., & Salguero-Gómez, R. (2020). Towards a
640 comparative framework of demographic resilience. *Trends in Ecology and*
641 *Evolution*, 35(9), 776–786.

642 Capdevila, P., Stott, I., Cant, J., Beger, M., Rowlands, G., Grace, M., & Salguero-
643 Gomez, R. (2021). Life history mediates the trade-offs among different
644 components of demographic resilience. *BioRxiv*.

645 Carmona, C. P., De Bello, F., Mason, N. W., & Lepš, J. (2016). Traits without borders:
646 Integrating functional diversity across scales. *Trends in Ecology and Evolution*,
647 31(5), 382–394.

648 Carnicer, J., Vives-Ingla, M., Blanquer, L., Méndez-Camps, X., Rosell, C., Sabaté, S.,
649 Gutiérrez, E., Sauras, T., Peñuelas, J., & Barbeta, A. (2021). Forest resilience

650 to global warming is strongly modulated by local-scale topographic,
651 microclimatic and biotic conditions. *Journal of Ecology*.

652 Carson, J. (1820). III. On the elasticity of the lungs. *Philosophical Transactions of the*
653 *Royal Society of London*, 110, 29–44.

654 Caswell, H. (1976). Community structure: A neutral model analysis. *Ecological*
655 *Monographs*, 46(3), 327–354.

656 Caswell, H. (2001). Matrix Population Models: Construction, Analysis, and
657 Interpretation. 2nd edn Sinauer Associates. Inc., Sunderland, MA.

658 CBD, U. (2010). Strategic plan for biodiversity 2011–2020 and the Aichi targets.
659 *Report of the Tenth Meeting of the Conference of the Parties to the Convention*
660 *on Biological Diversity*.

661 Ceballos, G., Ehrlich, P. R., Barnosky, A. D., García, A., Pringle, R. M., & Palmer, T.
662 M. (2015). Accelerated modern human-induced species losses: Entering the
663 sixth mass extinction. *Science Advances*, 1(5), e1400253.

664 Clark, A. T., Arnoldi, J.-F., Zelnik, Y. R., Barabas, G., Hodapp, D., Karakoç, C., König,
665 S., Radchuk, V., Donohue, I., & Huth, A. (2021). General statistical scaling laws
666 for stability in ecological systems. *Ecology Letters*, 24(7), 1474–1486.

667 Cole, L. E. S., Bhagwat, S. A., & Willis, K. J. (2014). Recovery and resilience of tropical
668 forests after disturbance. *Nature Communications*, 5.

669 Côté, I. M., Darling, E. S., & Brown, C. J. (2016). Interactions among ecosystem
670 stressors and their importance in conservation. *Proceedings of the Royal*
671 *Society B: Biological Sciences*, 283(1824), 20152592–20152592.

672 Dakos, V., Carpenter, S. R., van Nes, E. H., & Scheffer, M. (2014). Resilience
673 indicators: Prospects and limitations for early warnings of regime shifts.
674 *Philosophical Transactions of the Royal Society B: Biological Sciences*,
675 370(1659), 20130263–20130263.

676 Dale, M. R., & Fortin, M.-J. (2014). *Spatial analysis: A guide for ecologists*. Cambridge
677 University Press.

678 De Battisti, D. (2021). The resilience of coastal ecosystems: A functional trait-based
679 perspective. *Journal of Ecology*.

680 de Bello, F., Lavorel, S., Díaz, S., Harrington, R., Cornelissen, J. H., Bardgett, R. D.,
681 Berg, M. P., Cipriotti, P., Feld, C. K., & Hering, D. (2010). Towards an
682 assessment of multiple ecosystem processes and services via functional traits.
683 *Biodiversity and Conservation*, 19(10), 2873–2893.

684 de Vries, F. T., Liiri, M. E., Bjørnlund, L., Bowker, M. A., Christensen, S., Setälä, H.
685 M., & Bardgett, R. D. (2012). Land use alters the resistance and resilience of
686 soil food webs to drought. *Nature Climate Change*, 2(4), 276–280.

687 Delettre, O. (2021). Identity of ecological systems and the meaning of resilience.
688 *Journal of Ecology*.

689 Dennis, B., Ponciano, J. M., Lele, S. R., Taper, M. L., & Staples, D. F. (2006).
690 Estimating density dependence, process noise, and observation error.
691 *Ecological Monographs*, 76(3), 323–341.

692 DeSoto, L., Cailleret, M., Sterck, F., Jansen, S., Kramer, K., Robert, E. M., Aakala, T.,
693 Amoroso, M. M., Bigler, C., & Camarero, J. J. (2020). Low growth resilience to
694 drought is related to future mortality risk in trees. *Nature Communications*,
695 11(1), 1–9.

696 Díaz, S., Settele, J., Brondízio, E. S., Ngo, H. T., Agard, J., Arneth, A., Balvanera, P.,
697 Brauman, K. A., Butchart, S. H., & Chan, K. M. (2019). Pervasive human-driven
698 decline of life on Earth points to the need for transformative change. *Science*,
699 366(6471).

700 Dictionary, O. E. (1989). OED. 1989a), 5.

701 Donohue, I., Petchey, O. L., Kéfi, S., Génin, A., Jackson, A. L., Yang, Q., & O'Connor,
702 N. E. (2017). Loss of predator species, not intermediate consumers, triggers
703 rapid and dramatic extinction cascades. *Global Change Biology*, 23(8), 2962–
704 2972.

705 Donohue, I., Petchey, O. L., Montoya, J. M., Jackson, A. L., McNally, L., Viana, M.,
706 Healy, K., Lurgi, M., O'Connor, N. E., & Emmerson, M. C. (2013). On the
707 dimensionality of ecological stability. *Ecology Letters*, 16(4), 421–429.

708 Eagle, L. J. B., Milner, A. M., Klaar, M. J., Carrivick, J. L., Wilkes, M., & Brown, L. E.
709 (2021). Extreme flood disturbance effects on multiple dimensions of river
710 invertebrate community stability. *Journal of Animal Ecology*.

711 Folke, C., Carpenter, S., Walker, B., Scheffer, M., Elmqvist, T., Gunderson, L., &
712 Holling, C. S. (2004). Regime Shifts, Resilience, and Biodiversity in Ecosystem
713 Management. *Annual Review of Ecology, Evolution, and Systematics*, 35(1),
714 557–581.

715 Gallagher, R. V., Hughes, L., & Leishman, M. R. (2013). Species loss and gain in
716 communities under future climate change: Consequences for functional
717 diversity. *Ecography*, 36(5), 531–540.

718 Gladstone-Gallagher, R. V., Pilditch, C. A., Stephenson, F., & Thrush, S. F. (2019).
 719 Linking traits across ecological scales determines functional resilience. *Trends*
 720 *in Ecology and Evolution*, 34(12), 1080–1091.

721 Griffith, A. B., Salguero-Gómez, R., Merow, C., & McMahon, S. (2016). Demography
 722 beyond the population. *Journal of Ecology*, 104, 271–280.

723 Harrison, G. W. (1979). Stability under Environmental Stress: Resistance, Resilience,
 724 Persistence, and Variability. *The American Naturalist*, 113(5), 659–669.

725 Hodgson, D., McDonald, J. L., & Hosken, D. J. (2015). What do you mean, ‘resilient’?
 726 *Trends in Ecology and Evolution*, 30(9), 503–506.

727 Holling, C. S. (1973). Resilience and Stability of Ecological Systems. *Annual Review*
 728 *of Ecology and Systematics*, 4(1), 1–23.

729 Holling, C. S. (1996). Engineering resilience versus ecological resilience. *Engineering*
 730 *within Ecological Constraints*, 31(1996), 32.

731 Holling, C. S., & Goldberg, M. A. (1971). Ecology and planning. *Journal of the*
 732 *American Institute of Planners*, 37(4), 221–230.

733 Holling, C. S., & Orians, G. (1971). Toward an urban ecology. *Bulletin of the Ecological*
 734 *Society of America*, 52(2), 2–6.

735 Hoover, D. L., Pfennigwerth, A. A., & Duniway, M. C. (2021). Drought resistance and
 736 resilience: The role of soil moisture–plant interactions and legacies in a dryland
 737 ecosystem. *Journal of Ecology*.

738 Hughes, T. P., Baird, A. H., Bellwood, D. R., Card, M., Connolly, S. R., Folke, C.,
 739 Grosberg, R., Hoegh-Guldberg, O., Jackson, J. B., & Kleypas, J. (2003).

740 Climate change, human impacts, and the resilience of coral reefs. *Science*,
741 301(5635), 929–933.

742 Hughes, T. P., Linares, C., Dakos, V., van de Leemput, I. A., & van Nes, E. H. (2013).
743 Living dangerously on borrowed time during slow, unrecognized regime shifts.
744 *Trends in Ecology and Evolution*, 28(3), 149–155.

745 Ingrisch, J., & Bahn, M. (2018). Towards a comparable quantification of resilience.
746 *Trends in Ecology and Evolution*, 33(4), 251–259.

747 Ings, T. C., Montoya, J. M., Bascompte, J., Blüthgen, N., Brown, L., Dormann, C. F.,
748 Edwards, F., Figueroa, D., Jacob, U., & Jones, J. I. (2009). Ecological
749 networks—beyond food webs. *Journal of Animal Ecology*, 78(1), 253–269.

750 IPCC. (2021). *Climate Change 2021: The Physical Science Basis. Contribution of*
751 *Working Group I to the Sixth Assessment Report of the Intergovernmental*
752 *Panel on Climate Change [Masson-Delmotte, V., P. Zhai, A. Pirani, S. L.*
753 *Connors, C. Péan, S. Berger, N. Caud, Y. Chen, L. Goldfarb, M. I. Gomis, M.*
754 *Huang, K. Leitzell, E. Lonnoy, J. B. R. Matthews, T. K. Maycock, T. Waterfield,*
755 *O. Yelekçi, R. Yu and B. Zhou (eds.)].* Cambridge University Press.

756 Jentsch, A., & White, P. (2019). A theory of pulse dynamics and disturbance in
757 ecology. *Ecology*, 100, e02734.

758 Jia, Y., van der Heijden, M. G. A., Wagg, C., Feng, G., & Walder, F. (2020). Symbiotic
759 soil fungi enhance resistance and resilience of an experimental grassland to
760 drought and nitrogen deposition. *Journal of Ecology*.

761 Johnstone, J. F., Allen, C. D., Franklin, J. F., Frelich, L. E., Harvey, B. J., Higuera, P.
762 E., Mack, M. C., Meentemeyer, R. K., Metz, M. R., Perry, G. L., Schoennagel,
763 T., & Turner, M. G. (2016). Changing disturbance regimes, ecological memory,

764 and forest resilience. *Frontiers in Ecology and the Environment*, 14(7), 369–
765 378.

766 Jones, S. F., Stagg, C. L., Yando, E. S., James, W. R., Buffington, K. J., & Hester, M.
767 W. (2020). Stress gradients interact with disturbance to reveal alternative states
768 in salt marsh: Multivariate resilience at the landscape scale. *Journal of Ecology*.

769 Kattge, J., Bönisch, G., Díaz, S., Lavorel, S., Prentice, I. C., Leadley, P., Tautenhahn,
770 S., Werner, G. D., Aakala, T., & Abedi, M. (2020). TRY plant trait database–
771 enhanced coverage and open access. *Global Change Biology*, 26(1), 119–188.

772 Koricheva, J., & Gurevitch, J. (2014). Uses and misuses of meta-analysis in plant
773 ecology. *Journal of Ecology*, 102(4), 828–844.

774 Leclère, D., Obersteiner, M., Barrett, M., Butchart, S. H., Chaudhary, A., De Palma,
775 A., DeClerck, F. A., Di Marco, M., Doelman, J. C., & Dürauer, M. (2020).
776 Bending the curve of terrestrial biodiversity needs an integrated strategy.
777 *Nature*, 1–6.

778 Leizeaga, A., Hicks, L. C., Manoharan, L., Hawkes, C. V., & Rousk, J. (2020). Drought
779 legacy affects microbial community trait distributions related to moisture along
780 a savannah grassland precipitation gradient. *Journal of Ecology*.

781 Leverkus, A. B., Polo, I., Baudoux, C., Thorn, S., Gustafsson, L., & Rubio de Casas,
782 R. (2020). Resilience impacts of a secondary disturbance: Meta-analysis of
783 salvage logging effects on tree regeneration. *Journal of Ecology*.

784 Lewontin, R. C. (1969). The meaning of stability. *Brookhaven Symposia in Biology*,
785 22, 13–24.

786 Li, X., Yang, W., Gaedke, U., & de Ruiter, P. C. (2021). Energetic constraints imposed
787 on trophic interaction strengths enhance resilience in empirical and model food
788 webs. *Journal of Animal Ecology*.

789 Lipoma, M. L., Cabrol, D. A., Cuchiatti, A., Enrico, L., Gorné, L. D., & Díaz, S. (2021).
790 Low resilience at the early stages of recovery of the semi-arid Chaco forest—
791 Evidence from a field experiment. *Journal of Ecology*.

792 Lisovski, S., Gosbell, K., Minton, C., & Klaassen, M. (2020). Migration strategy as an
793 indicator of resilience to change in two shorebird species with contrasting
794 population trajectories. *Journal of Animal Ecology*.

795 Lloret, F., Keeling, E. G., & Sala, A. (2011). Components of tree resilience: Effects of
796 successive low-growth episodes in old ponderosa pine forests. *Oikos*, 120(12),
797 1909–1920.

798 MacArthur, R. (1955). Fluctuations of animal populations and a measure of community
799 stability. *Ecology*, 36(3), 533–536.

800 Magurran, A. E. (2013). *Measuring biological diversity*. John Wiley & Sons.

801 Maia, K. P., Marquitti, F. M. D., Vaughan, I. P., Memmott, J., & Raimundo, R. L. G.
802 (2021). Interaction generalisation and demographic feedbacks drive the
803 resilience of plant–insect networks to extinctions. *Journal of Animal Ecology*.

804 Matos, I. S., Menor, I. O., Rifai, S. W., & Rosado, B. H. P. (2020). Deciphering the
805 stability of grassland productivity in response to rainfall manipulation
806 experiments. *Global Ecology and Biogeography*, 29(3), 558–572.

807 Maxwell, S. L., Fuller, R. A., Brooks, T. M., & Watson, J. E. (2016). Biodiversity: The
808 ravages of guns, nets and bulldozers. *Nature News*, 536(7615), 143.

809 May, R. M. (1977). Thresholds and breakpoints in ecosystems with a multiplicity of
810 stable states. *Nature*, 269(5628), 471–477.

811 Medeiros, L. P., Song, C., & Saavedra, S. (2021). Merging dynamical and structural
812 indicators to measure resilience in multispecies systems. *Journal of Animal*
813 *Ecology*.

814 Miner, C. M., Burnaford, J. L., Ammann, K., Becker, B. H., Fradkin, S. C., Ostermann-
815 Kelm, S., Smith, J. R., Whitaker, S. G., & Raimondi, P. T. (2021). Latitudinal
816 variation in long-term stability of North American rocky intertidal communities.
817 *Journal of Animal Ecology*.

818 Moretti, M., Dias, A. T., De Bello, F., Altermatt, F., Chown, S. L., Azcarate, F. M., Bell,
819 J. R., Fournier, B., Hedde, M., & Hortal, J. (2017). Handbook of protocols for
820 standardized measurement of terrestrial invertebrate functional traits.
821 *Functional Ecology*, 31(3), 558–567.

822 Mougi, A., & Kondoh, M. (2012). Diversity of interaction types and ecological
823 community stability. *Science*, 337(6092), 349–351.

824 Mungi, N. A., Qureshi, Q., & Jhala, Y. V. (2021). Role of species richness and human-
825 impacts in resisting invasive species in tropical forests. *Journal of Ecology*.

826 Muñoz, R., Bongers, F., Rozendaal, D. M. A., González, E. J., Dupuy, J. M., & Meave,
827 J. A. (2021). Autogenic regulation and resilience in tropical dry forest. *Journal*
828 *of Ecology*.

829 Myhrvold, N. P., Baldridge, E., Chan, B., Sivam, D., Freeman, D. L., & Ernest, S. M.
830 (2015). An amniote life-history database to perform comparative analyses with
831 birds, mammals, and reptiles: Ecological Archives E096-269. *Ecology*, 96(11),
832 3109–3109.

833 Nelson, D., Busch, M. H., Kopp, D. A., & Allen, D. C. (2021). Energy pathways
834 modulate the resilience of stream invertebrate communities to drought. *Journal*
835 *of Animal Ecology*.

836 Neubert, M. G., & Caswell, H. (1997). Alternatives to Resilience for Measuring the
837 Responses of Ecological Systems to Perturbations. *Ecology*, 78(3), 653–665.

838 Newbold, T., Hudson, L. N., Hill, S. L., Contu, S., Lysenko, I., Senior, R. A., Börger,
839 L., Bennett, D. J., Choimes, A., & Collen, B. (2015). Global effects of land use
840 on local terrestrial biodiversity. *Nature*, 520(7545), 45–50.

841 Nowicki, R. J., Thomson, J. A., Fourqurean, J. W., Wirsing, A. J., & Heithaus, M. R.
842 (2021). Loss of predation risk from apex predators can exacerbate marine
843 tropicalization caused by extreme climatic events. *Journal of Animal Ecology*.

844 Oliver, T. H., Heard, M. S., Isaac, N. J. B., Roy, D. B., Procter, D., Eigenbrod, F.,
845 Freckleton, R., Hector, A., Orme, C. D. L., Petchey, O. L., Proença, V., Raffaelli,
846 D., Suttle, K. B., Mace, G. M., Martín-López, B., Woodcock, B. A., & Bullock, J.
847 M. (2015). Biodiversity and Resilience of Ecosystem Functions. *Trends in*
848 *Ecology and Evolution*, 30(11), 673–684.

849 Orr, J. A., Vinebrooke, R. D., Jackson, M. C., Kroeker, K. J., Kordas, R. L., Mantyka-
850 Pringle, C., Van den Brink, P. J., De Laender, F., Stoks, R., & Holmstrup, M.
851 (2020). Towards a unified study of multiple stressors: Divisions and common
852 goals across research disciplines. *Proceedings of the Royal Society B*,
853 287(1926), 20200421.

854 Ovenden, T. S., Perks, M. P., Clarke, T.-K., Mencuccini, M., & Jump, A. S. (2021). Life
855 after recovery: Increased resolution of forest resilience assessment sheds new

856 light on post-drought compensatory growth and recovery dynamics. *Journal of*
857 *Ecology*.

858 Paniw, M., de la Riva, E. G., & Lloret, F. (2021). Demographic traits improve
859 predictions of spatiotemporal changes in community resilience to drought.
860 *Journal of Ecology*.

861 Pecl, G. T., Araújo, M. B., Bell, J. D., Blanchard, J., Bonebrake, T. C., Chen, I.-C.,
862 Clark, T. D., Colwell, R. K., Danielsen, F., Evengård, B., Falconi, L., Ferrier, S.,
863 Frusher, S., Garcia, R. A., Griffis, R. B., Hobday, A. J., Janion-Scheepers, C.,
864 Jarzyna, M. A., Jennings, S., ... Williams, S. E. (2017). Biodiversity
865 redistribution under climate change: Impacts on ecosystems and human well-
866 being. *Science*, 355(6332), eaai9214–eaai9214.

867 Pérez-Valera, E., Verdú, M., Navarro-Cano, J. A., & Goberna, M. (2018). Resilience
868 to fire of phylogenetic diversity across biological domains. *Molecular Ecology*,
869 27(13), 2896–2908.

870 Pimm, S. L. (1984). The complexity and stability of ecosystems. *Nature*, 307(5949),
871 321.

872 Qian, J. J., & Akçay, E. (2020). The balance of interaction types determines the
873 assembly and stability of ecological communities. *Nature Ecology & Evolution*,
874 4(3), 356–365.

875 Ricklefs, R. E. (1977). On the evolution of reproductive strategies in birds:
876 Reproductive effort. *The American Naturalist*, 111(979), 453–478.

877 Rutter, M. (1979). Protective factors in children's responses to stress and
878 disadvantage. *Annals of the Academy of Medicine, Singapore*, 8(3), 324–338.

879 Scheffer, M., Bascompte, J., Brock, W. A., Brovkin, V., Carpenter, S. R., Dakos, V.,
880 Held, H., van Nes, E. H., Rietkerk, M., & Sugihara, G. (2009). Early-warning
881 signals for critical transitions. *Nature*, 461(7260), 53–59.

882 Scheffer, M., Bolhuis, J. E., Borsboom, D., Buchman, T. G., Gijzel, S. M. W., Goulson,
883 D., Kammenga, J. E., Kemp, B., Leemput, I. A. van de, Levin, S., Martin, C. M.,
884 Melis, R. J. F., Nes, E. H. van, Romero, L. M., & Rikkert, M. G. M. O. (2018).
885 Quantifying resilience of humans and other animals. *Proceedings of the*
886 *National Academy of Sciences*, 115(47), 11883–11890.

887 Scholes, R. J., Montanarella, L., Brainich, E., Barger, N., ten Brink, B., Cantele, M.,
888 Erasmus, B., Fisher, J., Gardner, T., & Holland, T. G. (2018). *IPBES (2018):*
889 *Summary for policymakers of the assessment report on land degradation and*
890 *restoration of the Intergovernmental Science-Policy Platform on Biodiversity*
891 *and Ecosystem Services*.

892 Serra, R. T., Santos, C. D., Rousseau, G. X., Triana, S. P., Muñoz Gutiérrez, J. A., &
893 Burgos Guerrero, J. E. (2021). Fast recovery of soil macrofauna in regenerating
894 forests of the Amazon. *Journal of Animal Ecology*.

895 Stearns, S. C. (1992). *The Evolution of Life Histories*. OUP Oxford.

896 Steel, Z., Foster, D., Lydersen, J., Stephens, Scott, Paudel, Asha, Markwith, Scott,
897 Merriam, Kyle, & Collins, Brandon M. (2021). Ecological resilience and
898 vegetation transition in the face of two successive large wildfires. *Journal of*
899 *Ecology*.

900 Stott, I., Townley, S., & Hodgson, D. J. (2011). A framework for studying transient
901 dynamics of population projection matrix models: A synthesis of transient
902 demography. *Ecology Letters*, 14(9), 959–970.

- 903 Su, H., Wang, R., Feng, Y., Li, Y., Li, Y., Chen, J., Xu, C., Wang, S., Fang, J., & Xie,
904 P. (2020). Long-term empirical evidence, early warning signals and multiple
905 drivers of regime shifts in a lake ecosystem. *Journal of Ecology*.
- 906 Sutherland, W. J., Freckleton, R. P., Godfray, H. C. J., Beissinger, S. R., Benton, T.,
907 Cameron, D. D., Carmel, Y., Coomes, D. A., Coulson, T., & Emmerson, M. C.
908 (2013). Identification of 100 fundamental ecological questions. *Journal of*
909 *Ecology*, 101(1), 58–67.
- 910 Thakur, M. P., van der Putten, W. H., Apon, F., Angelini, E., Vreš, B., & Geisen, S.
911 (2021). Resilience of rhizosphere microbial predators and their prey
912 communities after an extreme heat event. *Functional Ecology*, 35(1), 216–225.
- 913 Tredgold, T. (1818). XXXVII. On the transverse strength and resilience of timber. *The*
914 *Philosophical Magazine*, 51(239), 214–216.
- 915 Tsimara, E., Vasilakopoulos, P., Koutsidi, M., Raitzos, D. E., Lazaris, A., & Tzanatos,
916 E. (2021). An Integrated Traits Resilience Assessment of Mediterranean
917 fisheries landings. *Journal of Animal Ecology*.
- 918 Tylianakis, J. M., Didham, R. K., Bascompte, J., & Wardle, D. A. (2008). Global change
919 and species interactions in terrestrial ecosystems. *Ecology Letters*, 11(12),
920 1351–1363.
- 921 UNISDR, U. (2015). Sendai framework for disaster risk reduction 2015–2030.
922 *Proceedings of the 3rd United Nations World Conference on DRR, Sendai,*
923 *Japan*, 14–18.
- 924 United Nations General Assembly. (2015). *Transforming our world: The 2030 agenda*
925 *for sustainable development* (No. 9780874216561; pp. 1–5).

926 Van Meerbeek, K., Jucker, T., & Svenning, J.-C. (2021). Unifying the concepts of
 927 stability and resilience in ecology. *Journal of Ecology*.

928 Violle, C., Navas, M.-L., Vile, D., Kazakou, E., Fortunel, C., Hummel, I., & Garnier, E.
 929 (2007). Let the concept of trait be functional! *Oikos*, 116(5), 882–892.

930 Webster, C. L., Kilminster, K. L., Sánchez Alarcón, M., Bennett, K., Strydom, S.,
 931 McNamara, S., Lavery, P. S., & McMahon, K. M. (2021). Population-specific
 932 resilience of *Halophila ovalis* seagrass habitat to unseasonal rainfall, an
 933 extreme climate event in estuaries. *Journal of Ecology*.

934 Yang, Q., Fowler, M. S., Jackson, A. L., & Donohue, I. (2019). The predictability of
 935 ecological stability in a noisy world. *Nature Ecology & Evolution*, 3(2), 251–259.

936