



Research



Diversity-dependent effects probably influenced the diversification of species-rich crown ant subfamilies during the Cenozoic

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Author for correspondence:

Corentin Jouault

e-mail: corentin.jouault@oum.ox.ac.uk

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Corentin Jouault¹, Anshuman Swain^{2,3} and
Christine Sosiak⁴

¹Oxford University Museum of Natural History, University of Oxford, Parks Road, Oxford OX1 3PW, UK

²Department of Ecology and Evolutionary Biology, and ³Museum of Paleontology, University of Michigan, Ann Arbor, MI, USA

⁴Biodiversity and Biocomplexity Unit, Okinawa Institute of Science and Technology Graduate University, Kunigami District, Okinawa 904-0495, Japan

CJ, 0000-0002-3680-5172; AS, 0000-0002-9180-2222;
CS, 0000-0002-9057-8636

Disentangling the influence of biotic interactions from abiotic environmental changes is a central challenge in macroevolution. While ants' evolutionary history has been widely studied, often in relation to abiotic factors or plant associations, the role of intra- and inter-lineage interactions, whether competitive or facilitative, remains poorly understood. In this study, we use birth–death models within a Bayesian framework, integrating fossil and extant data, to investigate how interactions between the five most species-rich crown ant subfamilies may have shaped their diversification during the Cenozoic. Our results suggest that negative intraclade interactions within Dolichoderinae, Dorylinae, Formicinae and Ponerinae probably affected their diversification. We detect a signal of interactions between Formicinae and Myrmicinae and Dolichoderinae and Dorylinae, possibly reflecting long-term co-evolutionary dynamics. Notably, contrary to earlier hypotheses suggesting competition between Ponerinae and Myrmicinae, our results indicate a facilitative interaction between these two groups, but also between Formicinae and Dolichoderinae and Ponerinae and Myrmicinae, suggesting that coexistence sometimes promotes rather than inhibits diversification. Overall, our study provides the first assessment of diversity-dependent effects on the evolutionary history of ants,

establishing biotic interactions as a fundamental and quantifiable force in shaping macroevolutionary patterns of one of Earth's most successful animal groups.

1. Introduction

Over deep time, the diversity dynamics of lineages are shaped by biotic factors, such as changes in plant communities, resource competition and predation, and abiotic factors, including temperature and sea-level fluctuations [1–5]. This dichotomy is often conceptualized through two competing paradigms: the Red Queen (RQ) hypothesis, which posits that intrinsic biotic interactions are the primary engine of evolution, and the Court Jester (CJ) hypothesis, which argues that extrinsic abiotic perturbations drive major evolutionary shifts. While most research has traditionally emphasized the impact of abiotic changes on the rise and fall of clades (a focus on the CJ model), the role of biotic interactions (RQ dynamics), sometimes questioned, remains comparatively understudied, especially across large temporal and spatial scales, and using the fossil record [3,6–8]. However, recent macroevolutionary studies, drawing on a renewed interest in deep time ecological interactions within and between lineages, highlight the pivotal role of biotic factors in shaping diversification by affecting the rates of origination (i.e. speciation at levels higher than species) and extinction [4,9–12]. These studies have focused on quantifying the effects of two types of interactions: (i) diversity dependence, where changes in a clade's own diversity (i.e. accumulation of species) limit its origination and potentially increase its extinction rates; and (ii) interclade interactions, where changes in the diversity of one clade affect the diversification dynamics of other co-occurring clades [13–18].

Classically, diversity dependence has been attributed to a reduction in speciation rates or an increase in extinction rates as ecological niches become saturated [15,17,19,20]. This pattern is thought to result from the progressive filling of available ecological space, ultimately limiting opportunities for ecological speciation [15,21–25]. Such dynamics imply the existence of a clade-level carrying capacity, representing the maximum number of species that can coexist within a given lineage at a particular time. Interclade interactions encompass both negative interactions, such as competition (either by passive replacement and active displacement; [26]) or predation, which typically reduce origination and elevate extinction rates, and positive interactions, such as facilitation, where interactions between clades may enhance origination and buffer against extinction [4,10,27].

In deep time, both intraclade and interclade interactions can occur among taxa with similar ecology or under conditions of limited resources, particularly during or after key paleoevents (e.g. mass extinction events; ecosystem restructuring). These interactions may involve closely related species but can also arise between distantly related clades [4,12–14]. As a result, a lineage may diversify, decline or be replaced by another as ecological niches are created, vacated or filled [28,29]. While such dynamics have been mostly investigated in vertebrate groups, they remain poorly understood in invertebrates, particularly in insects. Yet insects represent the most diverse group of non-microbial organisms on Earth, with over one million described species [30]. Some lineages, such as ants (Formicidae), are not only ubiquitous but also function as keystone taxa and ecosystem engineers, contributing significantly to global biomass and influencing other species through biotic interactions [31,32].

With over 15 000 extant species, ants exhibit a remarkable range of ecological interactions, both competitive and cooperative, within and between lineages (figure 1) [33–36]. Negative interactions, potentially favouring extinction and limiting origination, are frequent among ants due to the tendency of many species to exploit similar ecological resources, such as nesting sites, prey or food sources [33,35,36]. This overlap often leads to intense competition within species or genera and among distantly related taxa [37,38]. These competitive dynamics give rise to diversity-dependent effects, which often manifest as constraints on spatial distribution, either during the establishment of new colonies (e.g. competition for nesting sites) or later in foraging areas (e.g. access to food resources such as aphids or extrafloral nectaries), and sometimes to the formation of well-defined colony territories [33,36]. These interactions have long been posited to act as mechanisms that shape ant diversification and community structure, even across deep time [11]. For instance, a recent study investigated potential diversity-dependent effects between stem and crown ants, examining their possible role in the extinction of stem ants during the Late Cretaceous, coinciding with and probably driven by the radiation of crown ants [11]. Similarly, researchers also hypothesized that during the Cenozoic, ponerine ants competed with myrmicines for prey and nesting sites, potentially constraining the diversification of ponerine ants [39]. From a macroevolutionary perspective, this would manifest as a

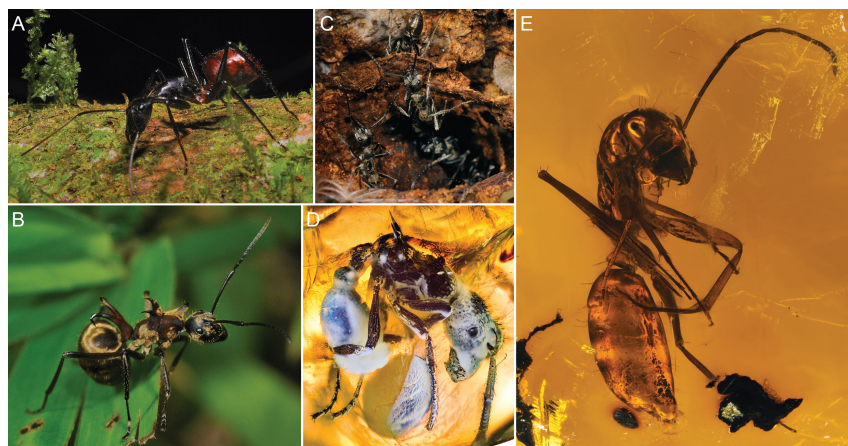


Figure 1. Ecological and morphological profiles of extant and extinct ants. (A) Formicinae: *Dinomyrmex gigas* (Latreille, 1802) (Sabah, Malaysia). (B) Formicinae: *Polyrhachis* sp. (Sabah, Malaysia). (C) Ponerinae: *Diacamma* sp. (Sabah, Malaysia). (D) extinct Myrmicinae (late Eocene, Baltic amber). (E) Formicinae: *Camponotus* sp. (mid-Miocene, Dominican amber). © Xavier Desmier (A–C); © Enrico Bonino (D,E).

decrease in origination rate and/or an increase in extinction rate of Ponerinae as Myrmicinae diversity increased—an expectation that can be explicitly tested using birth–death models [40]. It is also plausible that the diversification of generalist lineages occupying overlapping ecological niches—such as Formicinae and Dolichoderinae—led to negative interclade interactions. Similarly, the diversification of predatory lineages like Dorylinae may have negatively influenced the diversity trajectories of their prey lineages (e.g. Dolichoderinae). However, these hypotheses have yet to be empirically evaluated.

In contrast, complex mutualistic interactions (e.g. parabioses), potentially favouring origination and buffering against extinction, have been documented between certain ant species that share the same nest and may engage in interspecific trophallaxis or exploit common food resources [41]. At a more extreme level, some invasive ant species exhibit cooperative behaviours, which can enhance colony establishment and contribute to their dominance over native ant communities [42]. Such interactions are evident among the so-called ‘big five’ ant subfamilies (i.e. Myrmicinae, Formicinae, Ponerinae, Dolichoderinae and Dorylinae), most are generalist foragers (figure 1). However, if the evolutionary success of ants has long fascinated biologists [33,39], efforts to understand the drivers of this success in deep time have often focused on environmental factors, such as the diversification of flowering plants (e.g. [11,43]). In contrast, the potential role of diversity-dependent effects and interclade interactions within and between ant subfamilies has received comparatively little attention.

Using birth–death and diversity-dependent models within a Bayesian framework, and leveraging a combination of neontological and paleontological data, we aim to test for a macroevolutionary signal of Red Queen dynamics (i.e. competition and facilitation) among the five most speciose subfamilies of Formicidae—groups that together account for over 90% of ant diversity and possess the most complete fossil record. We investigated, *inter alia*, the long-standing hypothesis that links the diversity dynamics of ponerine ants to those of myrmicine ants [39]. Our analysis focused on the Cenozoic, a period marked by the rise and diversification of crown-group ants and dramatic abiotic change, providing a unique opportunity to assess the interplay of biotic and environmental forces [11,44,45].

2. Methods

2.1. Selection of ant subfamilies

Formicidae are highly diverse in modern ecosystems, but their species richness is unevenly distributed across subfamilies. Only five subfamilies are sufficiently speciose, in extant ecosystems (figure 2A) and the fossil record (figure 1D–E), to robustly investigate diversity-dependent effects: Dolichoderinae (approx. 5% of all extant species and approx. 5% of genera, 28 valid genera), Dorylinae (approx. 5% of species and approx. 8% genera, 27 valid genera), Formicinae (approx. 23% of species and approx. 16% of genera, 55 valid genera; figure 1A,B,E), Myrmicinae (approx. 50% of species and approx. 43% of genera, 148 valid genera; figure 1D) and Ponerinae (approx. 9% of species and approx. 14% of genera,

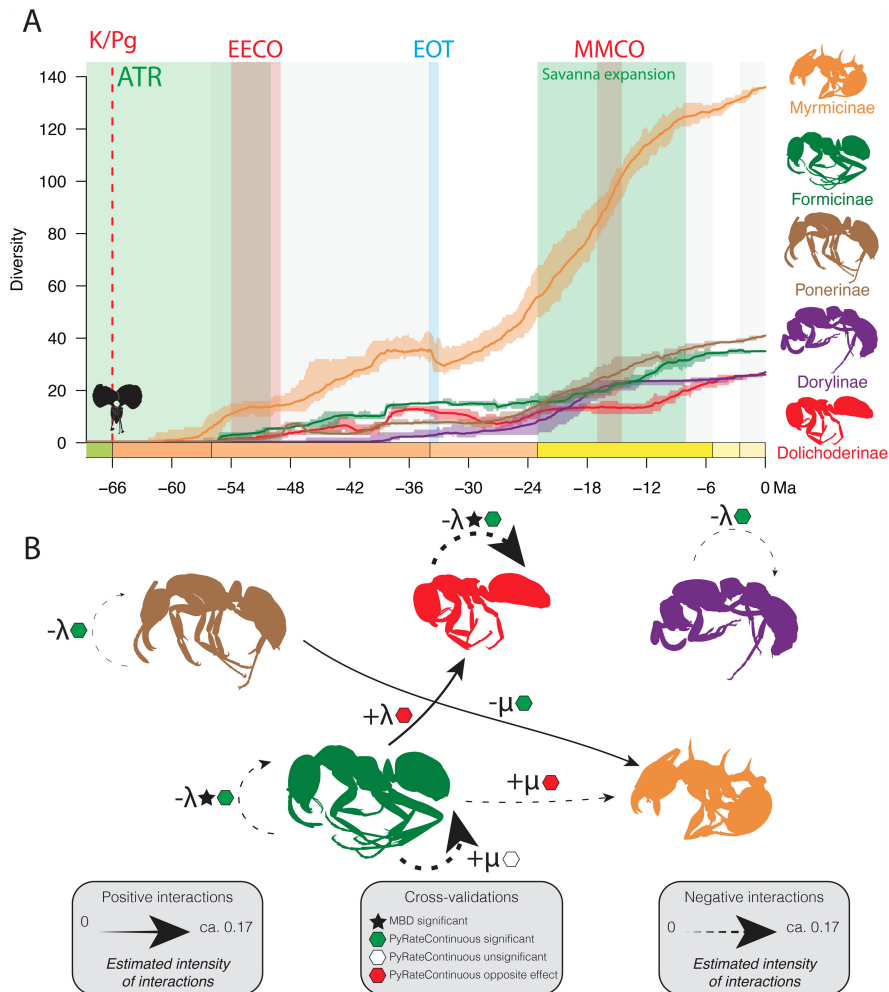


Figure 2. Diversity trajectories and the effect of diversity dependence or facilitation for the five most speciose ant subfamilies. (A) Diversity trajectories of five subfamilies, at the genus level, between the Cretaceous–Paleogene boundary and today. Reconstructions of diversity trajectories are from Jouault *et al.* [11] and have been replicated ten times, incorporating uncertainties around the ages of the fossil occurrences. For each plot, solid lines indicate mean posterior rates; shaded areas show 95% CI. (B) Network showing positive and negative interactions within and between ant subfamilies (only significant correlations are shown) reconstructed with the MCD model. Each arrow indicates the intensity of interaction imposed by a given guild towards another one. Colour of the geological timescale follows Cohen *et al.* [46] and abbreviations are as follows: ATR: angiosperm terrestrial revolution, EECO: early Eocene climatic optimum, EOT: Eocene–Oligocene transition, K/Pg: Cretaceous–Paleogene boundary, MMCO: middle Miocene climatic optimum. Silhouettes by Corentin Jouault.

50 valid genera; figure 1C) (based on AntWeb data) (figure 2A). Other subfamilies, such as Paraponerinae (monospecific in both extant and extinct ecosystems; approximately 0.01% of all extant species and approximately 0.29% of genera), were excluded from our analyses due to their comparatively low species richness and abundance in ecosystems.

The diversity dynamics of these ant lineages were recently quantified using a Bayesian framework that integrates both fossil and neontological data from time-calibrated phylogenies [11,47]. This approach has proven effective for reconstructing lineage diversification patterns by incorporating taxa not preserved in the fossil record and mitigating some of its inherent limitations [11,48]. Specifically, birth–death models were employed to estimate origination (i.e. genus-level speciation) and extinction rates over time, while accounting for sampling biases such as uneven preservation and the Lagerstätte effect. These models also incorporate time-variable preservation rates and accommodate uncertainties associated with the age of fossil occurrences [49–51].

We based our downstream analyses on the outputs from Jouault *et al.* [11], generated using reversible jump Markov chain Monte Carlo (RJMCMC), the recommended model for estimating origination and extinction rates with PyRate. These RJMCMC analyses integrated both palaeontological and neontological data, mitigating limitations associated with each dataset, such as distributional

bias or poor fossil representation of certain clades. The best-fitting preservation model was then selected after comparing non-homogeneous Poisson process, time-variable Poisson process and homogeneous Poisson process models, with a gamma model of rate heterogeneity incorporated to account for variation in preservation rates across lineages [11]. Origination (T_s) and extinction (T_e) times were averaged across 10 independent analyses to reduce the influence of outliers, and these estimates were used to generate lineage-through-time trajectories (all input files are available in electronic supplementary material, file S1). Overall, these combined precautions and approaches provide confidence in the long-term stability and reliability of the results.

All analyses were conducted at the genus level, a common practice in macroevolutionary research. This taxonomic scale offers several methodological advantages: (i) it is less sensitive to stratigraphic binning effects, since most fossil insect species are known from a single deposit or locality, whereas genera typically include multiple species and thus span broader temporal and spatial ranges; and (ii) it also tends to be more taxonomically stable than species [52,53].

Moreover, conducting analyses at the genus level helps reduce the influence of singletons, taxa represented by only one occurrence, which can otherwise inflate perceived diversification or interaction signals [50,51]. By aggregating species within genera, this approach diminishes noise from poorly sampled lineages and enhances the robustness of macroevolutionary inferences. While this necessarily smooths out species-level heterogeneity, it represents a reasonable and often necessary trade-off when dealing with the incompleteness of the fossil record. Furthermore, analyses performed at higher taxonomic levels, such as genus, are also particularly valuable for diminishing the prevalence of singletons because genera typically encompass multiple species.

In ants, the relative proportions of subfamilies within total extant diversity remain remarkably consistent when expressed as percentages of either genera or species (electronic supplementary material, file S1). The five most species-rich subfamilies are also the five richest in genera, indicating a stable, relatively scaling relationship across taxonomic levels. Consequently, although this assumption should be explicitly tested in future work, we expect that the broad macroevolutionary patterns identified at the genus level are probably representative of species-level trends as well.

2.2. Multiple clade diversity-dependence model analysis

To assess the potential effect of diversity-dependence on the diversity dynamics of the five largest ant subfamilies, we used the multiple clade diversity-dependence (MCDD) model in which origination and extinction rates are correlated with the diversity trajectory of other clades [4]. This model, implemented in PyRate [49], postulates that competitive interactions linked with an increase in diversity result in decreasing origination rates and/or increasing extinction rates. The MCDD model allows for testing diversity-dependence between taxa of a given clade and interclade interactions between taxa of distinct clades potentially sharing a similar ecology [4,10].

We extract the past diversity dynamics of the five ant subfamilies (Dolichoderinae: 45 genera; Dorylinae: 27 genera; Formicinae: 51 genera; Myrmicinae: 165 genera; and Ponerinae: 50 genera) using the number of living species at every point in time based on the T_s and T_e estimated under the RJMCMC model in Jouault *et al.* [40] (electronic supplementary material, file S1). We calculated ten diversity trajectories from the ten replicated analyses under the RJMCMC model. Because the fossil record of ants in the Mesozoic is much sparser than during the Cenozoic, and nearly exclusively composed of stem group lineages, we focused our study on the Cenozoic, which witnessed the rapid diversification of crown ants [11,44,47].

Our MCDD analyses were run and repeated on ten replicates (using the T_e and T_s estimated under the RJMCMC model) with 1 billion MCMC generations and a sampling frequency of 50 000. For each of the five ant subfamilies, we computed the median and the 95% HPD of the baseline origination and extinction rates (λ_i and μ_i), the within-group diversity-dependence parameters $g\lambda_i$ and $g\mu_i$, and the between-groups diversity dependence parameters $g\lambda_{ij}$ and $g\mu_{ij}$. In the MCDD model, the median of the sampled diversity dependence parameters $g\lambda_{ij} < 0$ and $g\mu_{ij} < 0$ indicates a positive interaction between clades, so that increasing diversity of a clade j correlates with higher speciation rates and lower extinction rates in clade i . Oppositely, $g\lambda_{ij} > 0$ and $g\mu_{ij} > 0$ indicate a negative interaction between clades, so that increasing diversity of a clade j correlates with lower speciation rates and higher extinction rates in clade i . Finally, $g\lambda_{ij} = 0$ and $g\mu_{ij} = 0$ imply that no diversity-dependent effects are detected and the diversification dynamics of clade i are independent of the diversity of clade j [8]. The interactions were considered significant when their median was different from 0 and the 95%

highest posterior density (HPD) intervals did not overlap with 0. Output files are available as electronic supplementary material, file S2, and a summary of the results in electronic supplementary material, table S1.

We monitored chain mixing and effective sample size (ESS) by examining the log files in Tracer 1.7.2 [54] and considered the convergence of parameters sufficient when their ESSs were greater than 200.

2.3. Multivariate birth–death model analyses

We also explored diversity-dependent effects using the multivariate birth–death (MBD) model [55]. This dual-model approach was chosen to ensure the robustness of our findings; the MCDD model serves as a powerful exploratory tool for detecting a complex web of interactions simultaneously, while the MBD model, which employs a gamma prior to control for over-parameterization, acts as a more conservative, corroborative test of the key interactions (see [8,55]).

The MBD model allows for changes in origination and extinction rates through time in relation to variables so that origination and extinction rates depend on the temporal variations of each factor. Therefore, it can be used to run a multiple clade diversity-dependence analysis by providing the diversity trajectories of each of the other four ant subfamilies, interacting with a given subfamily, as continuous variables. The diversity trajectories of ant subfamilies were generated by PyRate using the lineages-through-time generated by the RJMCMC analyses (*-lft* option).

An MCMC algorithm combined with a gamma prior (*-hsp 0*; recommended when testing only a few correlates), controlling for over-parameterization and for the potential effects of multiple testing, jointly estimates the baseline origination (λ_0) and extinction (μ_0) rates and all correlation parameters ($G\lambda$ and $G\mu$) [8]. In the MBD model, a correlation parameter is estimated to quantify the impact of each variable on origination and extinction independently. The strength and sign (positive or negative) of the correlations are estimated for each variable. When their 95% CI overlaps with zero and their shrinkage weight is below or close to 0.5, the correlation is considered non-significant. The results of the MBD analyses were summarized by calculating the posterior mean and 95% HDP of all correlation parameters and the mean of the respective shrinkage weights (across ten replicates), as well as the mean and 95% HDP of the baseline origination and extinction rates.

We ran the MBD model using 100 million MCMC generations and sampling every 50 000 to approximate the posterior distribution of all parameters (λ_0 , μ_0 , five $G\lambda$, five $G\mu$ and the shrinkage weights of each correlation parameter, ωG). We carried out five analyses, over the Cenozoic (*-maxT 66*), to estimate the effect of diversity-dependence between each of the five ant subfamilies of interest. We monitored chain mixing and ESS by examining the log files in Tracer 1.7.2 [54] and considered the convergence of parameters sufficient when their ESSs were greater than 200. Output files are provided as electronic supplementary material, file S2, and summaries of the results in electronic supplementary material, tables S2–S6.

2.4. Birth–death models with time-continuous correlates

To further assess the robustness of our results with respect to changes in the number and composition of clades included in the multivariate analyses, we employed a series of birth–death models with time-continuous correlates (*PyRateContinuous*) [49–51]. Compared with the multivariate MCDD and MBD frameworks, the *PyRateContinuous* analyses are inherently more conservative. *PyRateContinuous* fits a simplified birth–death process in which speciation and extinction rates are modelled separately as exponential functions of a single time-continuous covariate, using pre-estimated origination and extinction times (see §2.1) rather than the full occurrence data [49–51]. This greatly reduces the number of parameters and the amount of information available to detect subtle correlations. In addition, the use of an exponential link with broad, weakly informative priors on the correlation parameters ($G\lambda$ and $G\mu$) results in wider posterior distributions and significance is assessed strictly through the 95% HPD. Together, these factors cause weaker or marginal relationships identified in the more parameter-rich MCDD and MBD analyses to fall below the significance threshold in *PyRateContinuous*, making it a deliberately stringent tool for cross-validation rather than discovery.

We implemented these models in two complementary ways. First, we tested intraclade diversity dependence (option *-DD*) for the Dolichoderinae, Dorylinae, Formicinae, and Ponerinae, using an exponential model (*-m 0*), in which rates vary as exponential functions of the clades' own

diversity. Second, we tested interclade diversity dependence, evaluating pairwise relationships between subfamilies (e.g. Dolichoderinae and Dorylinae, Dolichoderinae and Formicinae) by fitting birth–death models where speciation and extinction rates vary through time as exponential functions ($-m\ 0$) of a time-continuous variable—specifically, the diversity of another clade.

In the first implementation, correlations were considered significant when the estimated diversity-dependence parameters ($G\lambda$ for speciation and $G\mu$ for extinction) were significantly different from zero based on their 95% HPD. The sign of the parameters determined the direction of the correlation (positive or negative, depending on whether $G \gg 0$ or $G \ll 0$). Because *PyRateContinuous* rescales covariates to range from 0 (present) to 1 (oldest), we inverted the covariate series so that larger values correspond to more recent times, ensuring that positive correlation coefficients ($G\lambda$, $G\mu$) indicate higher speciation or extinction rates with increasing diversity toward the present. The estimated correlation parameters ($G\lambda$ and $G\mu$) were considered statistically significant when their 95% HPD intervals did not overlap zero and were significantly different from zero, indicating a consistent directional association between the diversification dynamics of the focal and predictor clades.

All *PyRateContinuous* models were run for 50 million MCMC generations, with sampling every 50 000 generations. We assessed chain mixing and ESS using Tracer v.1.7.2 [54], and considered parameter convergence adequate when ESS values exceeded 200. Output files are provided as electronic supplementary material, file S2, and summary of the results in table 1.

3. Results

Our MCDD analysis reveals a series of diversity-dependent effects and inter-subfamily associations among ants (figure 2B; electronic supplementary material, table S1). Among the negative associations, we find that increases in diversity within several subfamilies are associated with decreases in their own origination rates. This is particularly evident in Dolichoderinae, where higher diversity correlates with reduced origination (median: $g\lambda = 0.1782$; 95% HPD interval: 0.0563–0.3). Similar trends are observed for Dorylinae ($g\lambda = 0.0374$; 95% HPD: 0–0.2833), Formicinae ($g\lambda = 0.048$; 95% HPD: 0–0.2072), and Ponerinae ($g\lambda = 0.0238$; 95% HPD: 0–0.2567), suggesting that higher subfamily-level diversity tends to co-occur with reduced origination within each group (figure 2B; electronic supplementary material, table S1).

Our results also indicate that increasing diversity within Formicinae is positively correlated with their own extinction rates (median: $g\mu = 0.1659$; 95% HPD: 0–0.2928), and correlates with a rise in extinction rates in Myrmicinae (median: $g\mu = 0.0483$; 95% HPD: 0–0.2776), pointing to possible inter-subfamily competitive dynamics (figure 2B; electronic supplementary material, table S1).

In contrast, we also detect positive correlations across subfamilies. Notably, increasing Formicinae diversity is associated with higher origination in Dolichoderinae (median: $g\lambda = -0.0927$; 95% HPD: -0.2466 to 0). Similarly, higher Ponerinae diversity co-occurs with reduced extinction rates in Myrmicinae (median: $g\mu = -0.0602$; 95% HPD: -0.2825 to 0), suggesting potential stabilizing or facilitative correlations between these clades (figure 2B; electronic supplementary material, table S1).

Using the more conservative MBD model (figure 3A–J), we also detected statistically significant diversity-dependent correlations in both Dolichoderinae (median: $G\lambda = -7.68$; 95% HPD: -14.6485 to -2.4175; figure 3A,B; electronic supplementary material, table S2) and Formicinae (median: $G\lambda = -8.1001$; 95% HPD: -17.5549 to -0.7087; figure 3E,F; electronic supplementary material, table S3), indicating that increasing diversity within these subfamilies is statistically associated with lower origination rates (figure 3).

Although the correlations detected in other clades were not statistically significant, our MBD results broadly support a consistent pattern of diversity-dependent dynamics across all subfamilies. Specifically, we observe a negative correlation between diversity and origination in Dorylinae (figure 3C,D; electronic supplementary material, table S4), a positive correlation between diversity and extinction in Formicinae (figure 3E,F; electronic supplementary material, table S3), and a negative correlation between diversity and origination in Ponerinae (figure 3I,J; electronic supplementary material, table S5).

In terms of interclade associations, the MBD model also identifies a statistically supported negative correlation between Dorylinae diversity and Dolichoderinae origination (median: $G\lambda = -0.2317$; 95% HPD: -0.4965 to -0.0021; figure 3A). Additionally, we recover similar (though not statistically significant) directional correlations as seen in our MCDD analysis with 1) an association between higher Formicinae diversity and increased origination in Dolichoderinae (figure 3A), 2) a correlation

Table 1. Results of PyRateContinuous analyses evaluating diversity-dependent relationships within and between pairs of clades (bold values indicate significant correlations).

		parameters	median	95% HPD	crossvalidations
diversity-dependent effects on Dolichoderinae	effect on origination (Dolichoderinae)	$G\lambda$	-1.9865	[-3.1279, -0.9967]	MCDD & MBD
	effect on origination (Dorylinae)	$G\lambda$	-1.0272	[-1.7434, -0.2816]	MBD
	effect on origination (Formicinae)	$G\lambda$	-1.7688	[-2.8449, -0.7429]	opposite MCDD
diversity-dependent effect on Dorylinae	effect on origination (Dorylinae)	$G\lambda$	-3.28	[-4.5457, -2.094]	MCDD
diversity-dependent effects on Formicinae	effect on origination (Formicinae)	$G\lambda$	-1.4853	[-2.3716, -0.4017]	MCDD & MBD
	effect on extinction (Formicinae)	$G\mu$	-1.0171	[-2.3385, 0.3042]	not significant
diversity-dependent effects on Myrmicinae	effect on extinction (Formicinae)	$G\mu$	-1.9158	[-3.3213, -0.7638]	opposite MCDD
	effect on extinction (Ponerinae)	$G\mu$	-2.235	[-3.4359, -1.1737]	MCDD
diversity-dependent effect on Ponerinae	effect on origination (Ponerinae)	$G\lambda$	-1.6274	[-2.5707, -0.7262]	MCDD

between higher Ponerinae diversity and reduced extinction in Myrmicinae (figure 3H), and 3) negative associations between Formicinae diversity and Myrmicinae extinction (figure 3H; electronic supplementary material, table S6).

Cross-validation of the correlations obtained with the MCDD and MBD models using the *PyRate-Continuous* birth–death framework provided significant statistical support for several intra-subfamily effects. Negative diversity-dependent effects on speciation were detected for Dolichoderinae (median: $G\lambda = -1.9865$; 95% HPD: -3.1279 to -0.9967), Dorylinae ($G\lambda = -3.28$; 95% HPD: -4.5457 to -2.094), Formicinae ($G\lambda = -1.4853$; 95% HPD: -2.3716 to -0.4017) and Ponerinae ($G\lambda = -1.6274$; 95% HPD: -2.5707 to -0.7262). In contrast, the intra-subfamily effect on Formicinae extinction was not statistically significant (table 1). Regarding inter-subfamily interactions, the effect of Dorylinae on Dolichoderinae origination displayed the same direction as the MBD model ($G\lambda = -1.0272$; 95% HPD: -1.7434 to -0.2816), and the effect of Ponerinae on Myrmicinae extinction the same direction as the MCDD model ($G\mu = -2.235$; 95% HPD: -3.4359 to -1.1737). The effect of Formicinae on Dolichoderinae origination ($G\lambda = -1.7688$; 95% HPD: -2.8449 to -0.7429), and Formicinae on Myrmicinae extinction ($G\mu = -1.9158$; 95% HPD: -3.3213 to -0.7638) are inverted when compared to the MCDD analyses.

4. Discussion

This study builds upon a recent broad-scale analysis of potential diversity-dependent effects in ants, which identified a significant signal when examining the entire evolutionary history of the clade based solely on palaeontological data (excluding singletons), thereby laying the groundwork for the present research [11]. However, that earlier study did not deeply investigate diversity-dependent interactions within or among crown ant lineages.

Expanding on these dynamics, our MCDD and MBD analyses reveal that most interclade effects are asymmetric, with significant correlations detected in only one direction (figures 2 and 3). We interpret this pattern as reflecting biological asymmetry, potentially arising from directional predation/competitive dominance or facilitative influences—such as those exerted by ecologically broader, more diverse or earlier-radiating clades on others. At local ecological scales, frameworks such as competition hierarchies (e.g. [56]) demonstrate that interspecific interactions are frequently unidirectional, with

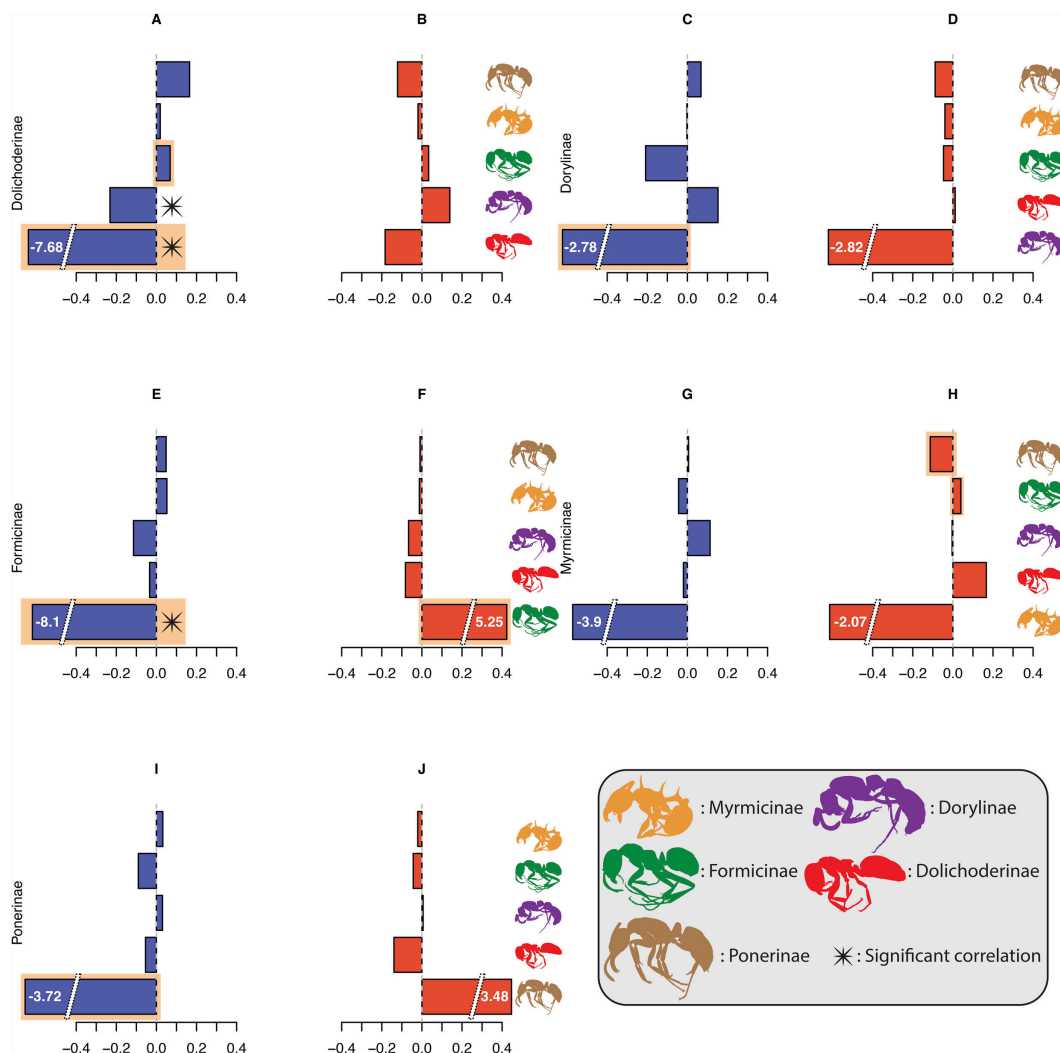


Figure 3. Correlations between changes in diversity of ant subfamilies and their diversification estimated with the MBD model. Bayesian inferences of correlation parameters on origination (blue: A,C,E,G,I) and extinction (red: B,D,F,H,J) with biotic factors. Stars indicate a significant correlation parameter for a given variable (shrinkage weights (w) > 0.5). The orange highlight shows correlations that were also recovered as statistically significant in the MCDD model analysis, underscoring the robustness of these findings. Silhouettes by Corentin Jouault.

dominant taxa suppressing subordinate ones. When these asymmetric dynamics are projected over geological timescales, they correspond to active displacement—a process in which the rise of one clade directly alters the diversification parameters of another, ultimately contributing to its decline [4,27,57,58]. Such asymmetry aligns with scenarios of predation, putative hierarchical competition, unidirectional diversity dependence, or facilitative interactions among ant lineages that we discuss below (see §4.2).

Our MBD analyses also tend to yield stronger correlation values for intraclade interactions than for interclade interactions (figures 2 and 3). Biologically, competitive interactions are generally expected to be stronger within subfamilies—where species often occupy overlapping ecological niches and compete directly for similar resources—than between subfamilies [36–38]. As a result, interclade effects may genuinely be smaller in magnitude.

Our *PyRateContinuous* analyses broadly corroborated the results obtained with the MBD and MCDD models, with two main exceptions: (i) the effect on extinction in Formicinae was not statistically significant, and (ii) the correlations between Formicinae and Dolichoderinae, as well as between Formicinae and Myrmicinae, exhibited opposite signs compared to those inferred from the MCDD analyses. *PyRateContinuous* evaluates each interclade relationship independently using rescaled time-continuous covariates, whereas MCDD estimates multiple interdependent correlations jointly within a hierarchical Bayesian framework, potentially capturing shared variance that reverses direction

when the relationship is tested alone. These structural differences can lead to divergences in both the direction and magnitude of inferred effects, potentially emphasizing different temporal components of the relationship, particularly for weak or borderline-significant interactions, which *in fine* reflects the greater sensitivity of MCDD to shared diversification dynamics and the more conservative, pairwise nature of *PyRateContinuous* analyses. The MBD analyses appear to support this hypothesis. As an intermediate model in terms of sensitivity, it reveals the same trends observed in the MCDD analyses, albeit without statistical significance. At present, it is not possible to fully rule out these two interactions, as they are also biologically plausible (see §4.2), but they will require further investigation in future studies.

4.1. Intraclade negative interactions in crown ants

Our MCDD analysis captures a pattern of intraclade negative interactions among Dolichoderinae, Ponerinae, Dorylinae and Formicinae (figures 2B and 3). These results are statistically supported by the more conservative MBD model for Dolichoderinae and Formicinae, and detected for Dorylinae and Ponerinae, but not statistically significant (figure 3A,C,E,I). In our analysis, Dolichoderinae and Formicinae exhibit the strongest intraclade negative effects on diversification ($g\lambda = 0.1782$ and $g\mu = 0.1659$, respectively). Notably, Formicinae show a unique pattern, with negative effects on both origination and extinction (statistically significant only for origination in our MBD analysis), whereas in other clades, the negative effect is observed only on origination (figures 2B and 3E,F). However, no significant intraclade negative effect was found for Myrmicinae, which are apparently more impacted by interclade interactions (figures 2B and 3G,H).

We propose that the stronger intraclade effects observed in Dolichoderinae and Formicinae can be explained by their ecological ubiquity and higher global abundance compared to Dorylinae or Ponerinae (e.g. [32,59–61]). This superabundance is particularly pronounced in arboreal communities. For instance, in the tropical lowland rainforests of the Peruvian Amazon, sampling over 62 000 ants revealed that a single Dolichoderinae species, *Dolichoderus bispinosus* (Olivier, 1792), dominated both the biomass (64.2%) and the individual counts (69%) [62]. A similar dominance pattern is observed in tropical Colombian lowland rainforests, where Dolichoderinae and Formicinae together represent approximately 60% of total ant abundance [60], and in Australian temperate rainforests as well [63,64]. Along altitudinal gradients in Europe or in local site studies in North America, Dolichoderinae and Formicinae likewise constitute the majority of both species diversity and abundance (e.g. [65,66]). Additionally, species within these subfamilies generally exhibit less specialized diets, which probably increases the probability of resource overlap and competition among closely related species. This notable absence of a negative signal in the hyper-diverse Myrmicinae suggests a distinct macroevolutionary dynamic that will be explored in detail below (see §4.4).

The unique pattern observed in Formicinae, characterized by both elevated extinction rates and slowed speciation rates as diversity increases, may be a result of markedly strong competition within certain lineages of this clade. *Formica* species in particular are noted to display high levels of intraspecific and interspecific aggression towards other members of the genus [67–69]. This genus is highly diverse, comprising around 180 extant species, and is hyper-abundant in most temperate ecosystems, where it often accounts for more than 50% of all collected specimens and has a biomass comparable to that of large mammalian predators [66,70]. They are also diverse and abundant in the fossil record, with 58 extinct species described to date, representing approximately 28% of all known extinct Formicinae (AntWeb). Battles between colonies occur frequently; during the spring, aggression is especially high and workers from other colonies comprise a significant proportion of a colony's diet [71]. Moreover, many *Formica* species are facultatively dulotic, relying on kidnapped workers from other congeneric colonies to favour their colony's survival and expansion [72]. Such behaviour directly reduces the targeted colony's fitness, and over macroevolutionary timescales, may be one of the plausible mechanisms contributing to elevated extinction rates within the subfamily. Additionally, in high northern latitudes, assemblages are frequently dominated by Formicinae, fostering interactions among species and increasing the likelihood of intraclade competition.

Evidence for intraclade negative interactions is observable in modern ecosystems and can provide insight into how ecological limits constrain diversification. For example, dolichoderine species of the ant genus *Azteca*, which commonly nest in *Cecropia* trees, often establish numerous incipient colonies on a single tree [73,74]. However, such coexistence is typically short-lived, culminating in intense intra- and interspecific competition. Ultimately, only one colony tends to dominate, eliminating others through direct fight [74,75]. This competition occurs both within species, where colonies race to grow

rapidly and monopolize space, and between closely related species, as multiple *Azteca* species may attempt to colonize the same tree simultaneously [73,74]. In the Formicinae subfamily, *Lasius niger* (Linnaeus, 1758) engages in pastoral behaviour, tending aphids to harvest honeydew. This activity often brings it into direct competition with other ant species, such as *Formica japonica* (Motschulsky, 1866). Conflicts frequently occur when these species encounter each other while tending aphids [76].

In predatory lineages such as the Dorylinae or some Ponerinae, the observed negative intraclade diversity dependence in diversification ($g\lambda = 0.0374$; $g\lambda = 0.0238$) probably reflects escalating competition over food resources as species richness increases. Within Dorylinae, particularly among army ants, many species exhibit varying degrees of dietary specialization, often preying predominantly on other ant species [77–79]. For example, studies in the La Selva Biological Station (Costa Rica) show that *Eciton burchellii* (Westwood, 1842) primarily targets *Camponotus*, *Neivamyrmex pilosus* (Smith, F., 1858) focuses on *Crematogaster*, and *Eciton dulcium* (Forel, 1912) preys on *Pachycondyla*. Other prey genera, such as *Pheidole*, are attacked by multiple species, including *Neivamyrmex gibbatus* Borgmeier, 1953, *Eciton vagans* (Olivier, 1792), and *Eciton mexicanum* Roger, 1863 [78], suggesting overlapping foraging niches. Despite some degree of ecological differentiation across prey identity, foraging strata (ground versus arboreal), and temporal activity (diurnal versus nocturnal), interspecific overlap remains common. When species such as *E. vagans* and *E. mexicanum* forage simultaneously in the same habitat and target the same prey (e.g. *Pheidole rugiceps* Wilson, 2003), direct competition becomes inevitable due to spatial and temporal limits on prey availability. While niche partitioning may buffer some interspecific conflict, it does not eliminate competition entirely and may instead concentrate competitive pressures among species [80–82]. Indeed, the significant dietary overlap on key, abundant prey resources like the genus *Pheidole* probably represents the specific ecological pressure point that manifests as a negative diversity-dependent signal over macroevolutionary time. Fossil evidence suggests that these behaviours have deep evolutionary origins, with signs of antagonistic interactions among stem ants, as evidenced by two species of *Gerontofornica* trapped while fighting in Kachin amber, dating back to the mid-Cretaceous [83].

While these are localized ecological interactions, they reflect a broader dynamic: as ant diversity increases, the intensity of competition for limited nesting sites (e.g. *Cecropia* tree) or ecological resources (e.g. prey or aphids) rises, reducing or constraining opportunities for the establishment and persistence of new lineages. Over evolutionary timescales, such niche saturation can slow diversification by decreasing speciation rates or increasing extinction risk, key features of diversity-dependent diversification. Similar mechanisms may operate across other taxa, especially those with tightly defined resource requirements and life-history constraints, illustrating how ecological limits can shape macroevolutionary patterns.

4.2. Interclade negative interactions in crown ants

Our results indicate that, over macroevolutionary timescales, intraclade interactions exert a stronger influence on diversification dynamics than interclade effects, which are often asymmetric. Among the eight significant diversity-dependent interactions we detected, five involved negative intraclade effects, while only one reflected a negative interclade interaction (figures 2B and 3). This pattern suggests that as diversity increases within a lineage, competition and other antagonistic interactions may sometimes increasingly constrain further diversification in another. Nonetheless, negative ecological interactions across taxonomic boundaries, whether between species, genera or subfamilies, remain frequent and ecologically consequential [84]. These include territorial aggression, chemical exclusion, competition for shared resources, predation and even social parasitism [33,85,86].

Interestingly, the asymmetry observed in our interclade effects, under both the MCDD and MBD models, probably reflects genuine ecological patterns consistent with the feeding behaviours of the lineages (figures 2 and 3). For instance, our MBD analysis reveals a significant negative correlation between the origination of Dolichoderinae and the increasing diversity of Dorylinae, potentially mirroring a unidirectional predator–prey dynamics (figure 3A). Dorylinae ants are predators that frequently target other ants, including Dolichoderinae, whereas the reverse scenario is less probable [78]. Thus, the directional interactions documented in modern ecosystems are plausibly echoed in our macroevolutionary results.

As highlighted in the *Azteca–Cecropia* case, competition among closely related ants for nesting sites and food resources is common. However, competition is not limited to close relatives; interclade competition also occurs when distantly related species exploit overlapping niches. For instance, *Cecropia* trees are inhabited not only by *Azteca*, but also by *Camponotus*, *Crematogaster*, *Neoponera*,

Pachycondyla and *Pseudomyrmex* [87], all of which compete for nesting space and access to glycogen-rich Müllerian bodies. The correlation observed in our *PyRateContinuous* analyses, in which increasing Formicinae diversity is associated with a reduction in Dolichoderinae origination rates, may reflect competitive interactions between these two generalist clades. Both share overlapping ecological niches and have a high abundance in past and present ecosystems (table 1). Our analyses further suggest that increasing diversity in Formicinae is associated with elevated extinction rates in Myrmicinae ($g\mu = 0.0483$), pointing to a long-term macroevolutionary signal of interclade antagonism, potentially driven by competition or even predation. Modern dominance hierarchies, generally driven by differences in foraging and aggression, are frequently dominated by formicines and dolichoderines, with myrmicines often being subordinate species. Such formicine-dominant hierarchies are especially pronounced in northern latitudes: taiga ant communities are dominated by *Formica* species, with *Myrmica* and *Leptothorax* being subordinate [88]; similarly, northern European island communities are dominated by *Formica* and *Lasius* species [89]. Interestingly, myrmicine species are more likely to be dominant in disturbed habitats today [36], making it possible that in the future the direction of this diversity-dependent interaction may switch. While our models do not detect the timing or biogeographic locality of diversity-dependent interactions, it is possible that across geological timescales, competitive suppression by dominant formicine species in formicine-rich latitudes could produce the macroevolutionary pattern of increased myrmicine extinction rates detected by our model (figures 2B and 3).

4.3. Facilitative interactions in crown ants

While our results highlight the significant role of negative interactions in shaping ant evolutionary history over time, they also reveal evidence of positive interactions between Formicinae and Dolichoderinae—also detected under the more conservative MBD model, but not statistically significant (figure 3A)—as well as between Ponerinae and Myrmicinae. Ants are well known not only for their competitive behaviours but also for their facilitative interactions, including complex forms of cooperation observed both within and between closely or distantly related species [33]. Our finding of a positive interaction whereby increased Ponerinae diversity is associated with reduced extinction in Myrmicinae is particularly noteworthy, as it runs counter to previous hypotheses of competition [39]. Although not statistically significant under the more conservative MBD model, the correlation was estimated in the same direction, supporting the robustness of this effect (figure 3H). A compelling ecological mechanism for this macroevolutionary pattern can be found in documented cases of commensalism or even parabiosis. For example, the tiny myrmicine ant *Strumigenys maynei* is known to form compound nests within the hollow-branch nests of the much larger ponerine ant *Platythyrea conradti* [90,91]. In this relationship, the *Strumigenys* workers scavenge on small prey thriving in the organic nest material collected by *P. conradti*, gaining a reliable food source and a protected environment [90,91], which would directly reduce their extinction risk. In this interaction, the ponerine host benefits from the nest defence capabilities of *S. maynei*. This specific, documented case of facilitation provides a powerful microevolutionary model for the large-scale pattern of reduced extinction detected by our analysis (figures 2B and 3). Moreover, such cases of facilitation also provide examples of effective niche ‘construction’, where the diversification of one clade leads to the creation of new microhabitats and niches for another, providing a mechanism that may buffer the latter against extinction and even drive its increased speciation. Ants are notable for shaping the biodiversity of many phylogenetically disparate taxa [31,92]; akin to myrmecophages and myrmecophiles, it is plausible that ecological diversification within a given clade leads to ecological diversification within others.

Another remarkable example of positive interactions among ants is found in ant–gardens, complex ant–plant mutualisms in which ants cultivate epiphytes in exchange for nesting space and nutritional rewards (e.g. [41,93–96]). These gardens typically involve species from the Formicinae, Dolichoderinae, Ponerinae and Myrmicinae subfamilies, though not all are necessarily present in every system. In Central and South America, ant–gardens are primarily maintained by genera such as *Crematogaster* and *Solenopsis* (Myrmicinae), *Azteca* and *Dolichoderus* (Dolichoderinae) and *Camponotus* (Formicinae), while *Anochetus* and *Odontomachus* (Ponerinae) are less commonly involved [97]. Within these ant–gardens, it is not uncommon to find two species exploiting the same food sources and cohabiting in non-obligatory, mixed colonies, where they tolerate each other and share a nest while maintaining separate brood chambers [98,99]. Both species actively defend their garden against a wide range of threats, including desiccation, invading ant species and phytophagous insects [100–104]. Such interactions can also reach a peak of complexity in mutualistic associations. For example, in the ant–gardens formed by

Crematogaster levior Longino, 2003 and *Camponotus femoratus* (Fabricius, 1804), the smaller *Crematogaster* species benefits from the nest-building activities of *Camponotus*, including seed-carrying behaviour and protection offered by the larger ants [94]. In return, *Camponotus* benefits from *Crematogaster*'s superior ability to locate food sources during foraging, which they often co-exploit or sometimes steal from the *Crematogaster* [94]. Our PyRateContinuous analyses, showing that increases in Formicinae diversity correlate with reduced extinction rates in Myrmicinae, may reflect such interactions, which could ultimately promote the persistence and survival of these clades (table 1). While this example focuses on interactions between formicine and myrmicine ants, similar cases of cohabitation and parabiosis have been documented between the Ponerinae (*Odontomachus panamensis* Forel, 1899 and *Neoponera goeldii* Forel, 1912) and the Myrmicinae (*Crematogaster limata parabiatica* Forel, 1904), with species from both subfamilies sharing the same nest and engaging in similar interactions [105]. Extended over long evolutionary timescales, these types of facilitative interactions offer a compelling ecological mechanism that could underlie the macroevolutionary pattern of increasing speciation or diminishing extinction rates detected by our models (figures 2B and 3).

4.4. The myrmicinae exception: diversification through niche partitioning and dispersion

The absence of a detectable negative intraclade signal in the Myrmicinae may represent a profound biological signal that speaks to the subfamily's exceptional evolutionary success [106]. As the most species-rich ant subfamily, accounting for roughly half of all described species (over 6000 extant species), Myrmicinae appear to have circumvented the diversity-dependent constraints that limit other clades. We propose that this is a direct consequence of their unparalleled adaptive radiation, which has facilitated niche partitioning and continuously expanded the available ecospace, but also the numerous dispersal events known for most of their genera [107–110].

Unlike the other subfamilies, which are sometimes more constrained in their ecology (e.g. predaceous Ponerinae ants), Myrmicinae have diversified along multiple niche axes (e.g. nesting strategies, feeding, colony size), thereby mitigating direct competition among their constituent lineages. First, while many myrmicines are generalists, the subfamily has produced remarkable trophic specialists. These include dedicated seed-harvesters in genera like *Pogonomyrmex* and *Messor* [33], and most famously, the tribe Attini, which evolved obligate fungus-farming—a unique agricultural strategy that created an entirely new trophic niche among ants [111,112]. Second, myrmicines occupy nearly every terrestrial habitat, from deep soil and leaf litter to the highest forest canopies [47,108,109,113,114]. This vertical and horizontal stratification allows different lineages to exploit different resources and avoid spatial overlap. Then, myrmicine colonies range from fewer than 50 individuals to massive supercolonies [115,116]. They exhibit a wide array of social structures (e.g. monogyny, polygyny), caste specializations (i.e. with or without dimorphism) and reproductive strategies (e.g. independent and dependent colony founding) [114,116–118]. This variation allows different species to adapt to different resource distributions and levels of habitat stability. Lastly, myrmicine ants possess an exceptional capacity for dispersal, achieved through a variety of strategies including long-distance nuptial flights, rafting across bodies of water and colony budding, each enabling the colonization of habitats far beyond their initial or native range [107,109,114,119]. This high dispersal ability probably facilitated the expansion of populations into novel or underexploited ecosystems rather than concentrating diversification within already saturated habitats. By continually accessing new ecological opportunities, myrmicines can circumvent the density-dependent constraints on speciation that often arise from competition in established communities. In other words, rather than intensifying resource overlap and competitive exclusion within the clade, dispersal-driven range expansion diffuses potential conflicts by spreading lineages across a broader array of environmental contexts. Over evolutionary time, this dynamic may contribute to both the clade's remarkable global distribution and the relatively weak signature of intraspecific diversity dependence detected in our analyses (figures 2B and 3).

Therefore, the absence of a negative intraclade signal is, paradoxically, strong evidence for the power of the diversity-dependence hypothesis. The model detects the 'filling' of available ecospace. For most subfamilies, this ecospace is a relatively contained niche that eventually becomes saturated. For Myrmicinae, however, constant dispersion and innovation have created a vast and expanding mosaic of semi-independent niches. A fungus-farming attine does not compete with a seed-harvesting *Messor*. The model correctly detects that, at the broad subfamily level, this mosaic is not 'full,' transforming a puzzling result into a powerful conclusion about the nature of adaptive radiation.

4.5. Future work

Our analyses are independent of phylogenetic framework, although they incorporate information derived from time-calibrated phylogenies, due to the nature of the models employed. Nonetheless, diversity-dependence models designed to detect similar effects using time-calibrated phylogenetic trees have been developed for over a decade. Early versions of these models were relatively simplistic, notably due to their assumption of zero extinction [15,120], but have since been refined to incorporate more realistic dynamics, including non-zero extinction rates [18,121].

While these methods have been predominantly applied to vertebrate or plant clades, they are equally applicable, and arguably underutilized, in the study of invertebrate evolution [15,122]. Specifically, exponential diversity-dependence (DDX) or diversity-dependence (DDL) models, either with or without a parameter model for extinction (e.g. DDL-E and DDL+E), could provide a powerful phylogenetic test of our findings. The main limitation to their application lies in the difficulty of achieving comprehensive sampling for extant species, particularly within highly diverse insect groups that can contain thousands of species. However, ants represent an ‘exception’ [121]. Thanks to the availability of molecular data, a significant proportion of ant diversity can be included in time-calibrated phylogenies [47], making them suitable candidates for such tree-based approaches. We therefore advocate for the development of a phylogeny-based counterpart to our study, which would allow for cross-validation of the trends and correlations identified through ‘fossil-based’ inferences.

Our study focuses on the most speciose ant subfamilies, which also possess the most temporally complete fossil records during the Cenozoic. However, other subfamilies that are ecologically significant in modern ecosystems, particularly in tropical regions, such as Ectatomminae, were not included in our analysis. This exclusion is due to the current limitations of their fossil record, which we consider insufficient to support robust temporal and statistical inferences [123]. Looking ahead, we hope that future discoveries and improvements in the fossil record of these subfamilies will enable the refinement of our hypotheses. In particular, efforts to redescribe and describe promising Cenozoic paleomyrmecofaunas (e.g. Fushun or Oise amber faunas) may help expand our understanding of their macroevolutionary dynamics.

5. Conclusion

Our findings demonstrate that the diversification of ants during the Cenozoic was not driven by biotic or abiotic factors in isolation, but by their continuous interplay. The persistent signals of competition and facilitation we detect are clear evidence of Red Queen dynamics, where the evolutionary fate of each subfamily was intrinsically linked to the diversity of its relatives and ecological counterparts. However, these biotic interactions played out on an environmental ‘stage’ that was constantly being reshaped by the Court Jester’s abiotic forces.

The major paleoevents of the Cenozoic (figure 2A) probably modulated the intensity and nature of these interactions. For example, the global cooling and aridification associated with the Eocene–Oligocene transition drove the expansion of grasslands and savannas. This biome turnover would have created vast new ecological opportunities, potentially favouring generalist foragers common in Formicinae and Myrmicinae and thereby intensifying the competitive interactions our model detects between them. In this view, abiotic events (the Court Jester) periodically altered the ecological ‘rules of the game,’ which in turn altered the landscape of competition and facilitation (the Red Queen) that proximately shaped diversification rates. This study thus provides a compelling case study of how continuous Red Queen processes act as the engine of diversification, while the track is laid and periodically re-routed by the forces of the Court Jester.

Ethics. This work did not require ethical approval from a human subject or animal welfare committee.

Data accessibility. The original dataset is extracted from Jouault *et al.* [11] and available in the following Figshare digital repository: <https://figshare.com/s/1d6a5e5639e797370c05> or as supporting information in the original paper. All the supplementary files used for the analyses are also available in the electronic supplementary material [124].

Supplementary material is available online [125].

Declaration of AI use. We have not used AI-assisted technologies in creating this article.

Authors’ contributions. C.J.: conceptualization, data curation, formal analysis, investigation, methodology, software, validation, visualization, writing—original draft, writing—review and editing; A.S.: investigation, validation, writing—original draft, writing—review and editing; C.S.: conceptualization, investigation, validation, writing—original draft, writing—review and editing.

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References

- Huxley JS. 1942 *Evolution: the modern synthesis*. New York, NY: Harper Press.
- Simpson GG. 1944 *Tempo and mode in evolution*. New York, NY: Columbia University Press. (doi:10.7312/simp93040)
- Benton MJ. 2009 The red queen and the court jester: species diversity and the role of biotic and abiotic factors through time. *Science* **323**, 728–732. (doi:10.1126/science.1157719)
- Silvestro D, Antonelli A, Salamin N, Quental TB. 2015 The role of clade competition in the diversification of North American canids. *Proc. Natl Acad. Sci. USA* **112**, 8684–8689. (doi:10.1073/pnas.1502803112)
- Condamine FL, Rolland J, Morlon H. 2019 Assessing the causes of diversification slowdowns: temperature-dependent and diversity-dependent models receive equivalent support. *Ecol. Lett.* **22**, 1900–1912. (doi:10.1111/ele.13382)
- Stanley SM. 1973 Effects of competition on rates of evolution, with special reference to bivalve mollusks and mammals. *Syst. Zool.* **22**, 486–506. (doi:10.2307/2412955)
- Benton MJ. 1996 Testing the roles of competition and expansion in tetrapod evolution. *Proc. R. Soc. B* **263**, 641–646. (doi:10.1098/rspb.1996.0096)
- Silvestro D, Pires MM, Quental TB, Salamin N. 2017 Bayesian estimation of multiple clade competition from fossil data. *Evol. Ecol. Res.* **18**, 41–59.
- Castiglione S, Mondanaro A, Carotenuto F, Passaro F, Fortelius M, Raia P. 2017 The many shapes of diversity: ecological and evolutionary determinants of biodiversity through time. *Evol. Ecol. Res.* **18**, 25–39.
- Jouault C, Nel A, Perrichot V, Legendre F, Condamine FL. 2022 Multiple drivers and lineage-specific insect extinctions during the Permo-Triassic. *Nat. Commun.* **13**, 7512. (doi:10.1038/s41467-022-35284-4)
- Jouault C, Condamine FL, Legendre F, Perrichot V. 2024 The angiosperm terrestrial revolution buffered ants against extinction. *Proc. Natl Acad. Sci. USA* **121**, e2317795121. (doi:10.1073/pnas.2317795121)
- Boderau M, Nel A, Jouault C. 2025 Diversification and extinction of Hemiptera in deep time. *Commun. Biol.* **8**, 352. (doi:10.1038/s42003-025-07773-x)
- Valentine JW. 1973 *Evolutionary paleoecology of the marine biosphere*. Englewood Cliffs, NJ: Prentice-Hall.
- Sepkoski JJ. 1978 A kinetic model of phanerozoic taxonomic diversity. I. Analysis of marine orders. *Paleobiology* **4**, 223–251. (doi:10.1017/S0094837300005972)
- Rabosky DL, Lovette IJ. 2008 Density-dependent diversification in North American wood warblers. *Proc. R. Soc. B Biol. Sci.* **275**, 2363–2371. (doi:10.1098/rspb.2008.0630)
- Gavrilets S, Losos JB. 2009 Adaptive radiation: contrasting theory with data. *Science* **323**, 732–737. (doi:10.1126/science.1157966)
- Quental TB, Marshall CR. 2009 Extinction during evolutionary radiations: reconciling the fossil record with molecular phylogenies. *Evolution* **63**, 3158–3167. (doi:10.1111/j.1558-5646.2009.00794.x)
- Etienne RS, Haegeman B. 2012 A conceptual and statistical framework for adaptive radiations with a key role for diversity dependence. *Am. Nat.* **180**, E75–E89. (doi:10.1086/667574)
- Ricklefs RE, Cox GW. 1972 Taxon cycles in the West Indian Avifauna. *Am. Nat.* **106**, 195–219. (doi:10.1086/282762)
- Nee S, Mooers AO, Harvey PH. 1992 Tempo and mode of evolution revealed from molecular phylogenies. *Proc. Natl Acad. Sci. USA* **89**, 8322–8326. (doi:10.1073/pnas.89.17.8322)
- Simpson GG. 1953 *The major features of evolution*. New York, NY: Columbia University Press. (doi:10.7312/simp93764)
- Valentine JW. 1980 Determinants of diversity in higher taxonomic categories. *Paleobiology* **6**, 444–450. (doi:10.1017/s0094837300003614)
- Valentine JW. 1985 Biotic diversity and clade diversity. In *Panerozoic diversity patterns* (ed. JW Valentine), pp. 419–424. Princeton, NJ: Princeton University Press. (doi:10.1515/9781400855056.419)
- Walker TD, Valentine JW. 1984 Equilibrium models of evolutionary species diversity and the number of empty niches. *Am. Nat.* **124**, 887–899. (doi:10.1086/284322)

25. Schluter D. 2000 *The ecology of adaptive radiation*, pp. 236–243. Oxford, UK: Oxford University Press. (doi:10.1093/oso/9780198505235.003.0010)
26. Sepkoski JJ. 1996 Competition in macroevolution: the double wedge revisited. In *Evolutionary paleobiology* (eds D Jablonski, DH Erwin, JH Lipps), pp. 211–255. Chicago, IL: University of Chicago Press.
27. Condamine FL, Romieu J, Guinot G. 2019 Climate cooling and clade competition likely drove the decline of lamniform sharks. *Proc. Natl Acad. Sci. USA* **116**, 20584–20590. (doi:10.1073/pnas.1902693116)
28. Liow LH, Reitan T, Harnik PG. 2015 Ecological interactions on macroevolutionary time scales: clams and brachiopods are more than ships that pass in the night. *Ecol. Lett.* **18**, 1030–1039. (doi:10.1111/ele.12485)
29. Pires MM, Silvestro D, Quental TB. 2017 Interactions within and between clades shaped the diversification of terrestrial carnivores. *Evolution* **71**, 1855–1864. (doi:10.1111/evo.13269)
30. Stork NE. 2018 How many species of insects and other terrestrial arthropods are there on earth? *Annu. Rev. Entomol.* **63**, 31–45. (doi:10.1146/annurev-ento-020117-043348)
31. Parker J, Kronauer DJC. 2021 How ants shape biodiversity. *Curr. Biol.* **31**, R1208–R1214. (doi:10.1016/j.cub.2021.08.015)
32. Schultheiss P, Nooten SS, Wang R, Wong MKL, Brassard F, Guénard B. 2022 The abundance, biomass, and distribution of ants on Earth. *Proc. Natl Acad. Sci. USA* **119**, e2201550119. (doi:10.1073/pnas.2201550119)
33. Hölldobler B, Wilson EO. 1990 *The ants*. Cambridge, MA: The Belknap Press of Harvard University Press.
34. California Academy of Science. Antweb version 8.114. See <https://www.antweb.org> (accessed May 2025).
35. Lach L, Parr CL, Abbott K. 2010 *Ant ecology*. Oxford, UK: Oxford University Press. (doi:10.1093/acprof:oso/9780199544639.001.0001)
36. Cerdá X, Arnan X, Retana J. 2013 Is competition a significant hallmark of ant (Hymenoptera: Formicidae) ecology? *Myrmecol. News* **18**, 131–147. (doi:10.25849/myrmecol.news_018:131)
37. Davidson DW, Cook SC, Snelling RR, Chua TH. 2003 Explaining the abundance of ants in lowland tropical rainforest canopies. *Science* **300**, 969–972. (doi:10.1126/science.1082074)
38. Mottl O, Yombai J, Novotný V, Leponce M, Weiblen GD, Klimeš P. 2021 Inter-specific aggression generates ant mosaics in canopies of primary tropical rainforest. *Oikos* **130**, 1087–1099. (doi:10.1111/oik.08069)
39. Wilson EO, Hölldobler B. 2005 The rise of the ants: a phylogenetic and ecological explanation. *Proc. Natl Acad. Sci. USA* **102**, 7411–7414. (doi:10.1073/pnas.0502264102)
40. Jouault C, Condamine FL, Legendre F, Perrichot V. 2024 Reply to vermeij: challenges and opportunities in macroevolution. *Proc. Natl Acad. Sci. USA* **121**, e2408795121. (doi:10.1073/pnas.2408795121)
41. Orivel J, Errard C, Dejean A. 1997 Ant gardens: interspecific recognition in parabiocytic ant species. *Behav. Ecol. Sociobiol.* **40**, 87–93. (doi:10.1007/s002650050319)
42. Angulo E *et al.* 2024 The Argentine ant, *Linepithema humile*: natural history, ecology and impact of a successful invader. *Entomol. Gen.* **44**, 41–61. (doi:10.1127/entomologia/2023/2187)
43. Moreau CS, Bell CD, Vila R, Archibald SB, Pierce NE. 2006 Phylogeny of the ants: diversification in the age of angiosperms. *Science* **312**, 101–104. (doi:10.1126/science.1124891)
44. Sosiak C, Cockx P, Suarez PA, McKellar R, Barden P. 2024 Prolonged faunal turnover in earliest ants revealed by North American Cretaceous amber. *Curr. Biol.* **34**, 1755–1761. (doi:10.1016/j.cub.2024.02.058)
45. Borowiec ML, Zhang YM, Neves K, Ramalho MO, Fisher BL, Lucky A, Moreau CS. 2025 Evaluating UCE data adequacy and integrating uncertainty in a comprehensive phylogeny of ants. *Syst. Biol.* **74**, 700–722. (doi:10.1093/sysbio/syaf001)
46. Cohen KM, Finney SC, Gibbard PL, Fan JX. 2024 The ICS international chronostratigraphic chart. *Episodes* **36**, 199–204. (doi:10.18814/epiugs/2013/v36i3/002)
47. Economo EP, Narula N, Friedman NR, Weiser MD, Guénard B. 2018 Macroecology and macroevolution of the latitudinal diversity gradient in ants. *Nat. Commun.* **9**, 1778. (doi:10.1038/s41467-018-04218-4)
48. Brée B, Condamine FL, Guinot G. 2022 Combining palaeontological and neontological data shows a delayed diversification burst of carcharhiniform sharks likely mediated by environmental change. *Sci. Rep.* **12**, 21906. (doi:10.1038/s41598-022-26010-7)
49. Silvestro D, Salamin N, Schnitzler J. 2014 PyRate: a new program to estimate speciation and extinction rates from incomplete fossil data. *Methods Ecol. Evol.* **5**, 1126–1131. (doi:10.1111/2041-210X.12263)
50. Silvestro D, Schnitzler J, Liow LH, Antonelli A, Salamin N. 2014 Bayesian estimation of speciation and extinction from incomplete fossil occurrence data. *Syst. Biol.* **63**, 349–367. (doi:10.1093/sysbio/syu006)
51. Silvestro D, Salamin N, Antonelli A, Meyer X. 2019 Improved estimation of macroevolutionary rates from fossil data using a Bayesian framework. *Paleobiology* **45**, 546–570. (doi:10.1017/pab.2019.23)
52. Allmon WD. 1992 Genera in paleontology: definition and significance. *Hist. Biol.* **6**, 149–158. (doi:10.1080/10292389209380424)
53. Foote M. 2000 Origination and extinction components of taxonomic diversity: general problems. *Paleobiology* **26**, 74–102. (doi:10.1017/S0094837300026890)
54. Rambaut A, Drummond AJ, Xie D, Baele G, Suchard MA. 2018 Posterior summarization in Bayesian phylogenetics using tracer 1.7. *Syst. Biol.* **67**, 901–904. (doi:10.1093/sysbio/syy032)
55. Lehtonen S, Silvestro D, Karger DN, Scotese C, Tuomisto H, Kessler M, Peña C, Wahlberg N, Antonelli A. 2017 Environmentally driven extinction and opportunistic origination explain fern diversification patterns. *Sci. Rep.* **7**, 4831. (doi:10.1038/s41598-017-05263-7)

56. Savolainen R, Vepsäläinen K. 1988 A competition hierarchy among boreal ants: impact on resource partitioning and community structure. *Oikos* **51**, 135–155. (doi:10.2307/3565636)
57. Stanley SM, Newman WA. 1980 Competitive exclusion in evolutionary time: the case of the acorn barnacles. *Paleobiology* **6**, 173–183. (doi:10.1017/S0094837300006734)
58. Lidgard S, Di Martino E, Zágóršek K, Liow LH. 2021 When fossil clades ‘compete’: local dominance, global diversification dynamics and causation. *Proc. R. Soc. B Biol. Sci.* **288**, 20211632. (doi:10.1098/rspb.2021.1632)
59. Adis J, Harada AY, Fonseca C da, Paarmann W, Rafael JA. 1998 Arthropods obtained from the Amazonian tree species ‘Cupiuba’ (*Goupia glabra*) by repeated canopy fogging with natural pyrethrum. *Acta Amazon.* **28**, 273–273. (doi:10.1590/1809-43921998283283)
60. Chacón de Ulloa P, Valdés-Rodríguez S, Hurtado-Giraldo A, Pimienta MC. 2014 Hormigas arbóreas del parque nacional natural gorgona (Pacífico de Colombia). *Rev. De Biol. Trop.* **62**, 277–287. (doi:10.15517/rbt.v62i0.16341)
61. Guénard B, Weiser M, Gomez K, Narula N, Economo EP. 2017 The global ant biodiversity informatics (GABI) database: a synthesis of ant species geographic distributions. *Myrmecol. News* **24**, 83–89. (doi:10.25849/myrmecol.news_024:083)
62. Tobin JE. 1997 Competition and coexistence of ants in a small patch of rainforest canopy in Peruvian Amazonia. *J. N. Y. Entomol. Soc.* **105**, 105–112.
63. Majer JD. 1990 The abundance and diversity of arboreal ants in Northern Australia. *Biotropica* **22**, 191–199. (doi:10.2307/2388412)
64. Majer JD, Kitching RL, Heterick BE, Hurley K, Brennan KEC. 2001 North–South patterns within arboreal ant assemblages from rain forests in eastern Australia. *Biotropica* **33**, 643–661. (doi:10.1111/j.1744-7429.2001.tb00222.x)
65. Uno S, Cotton J, Philpott SM. 2010 Diversity, abundance, and species composition of ants in urban green spaces. *Urban Ecosyst.* **13**, 425–441. (doi:10.1007/s11252-010-0136-5)
66. Schifani E, Grasso DA, Gobbi M, Spotti FA, Pedrotti L, Vettorazzo E, Mori A, Castracani C. 2024 Ant diversity along elevational gradients in the European alps: insights for conservation under a changing climate. *J. Insect Conserv.* **28**, 401–413. (doi:10.1007/s10841-023-00546-z)
67. Moli FL, Parmigiani S. 1982 Intraspecific combat in the red wood ant (*Formica lugubris*, Zett.). *Aggress. Behav.* **8**, 145–148. (doi:10.1002/1098-2337(1982)8:2<145::AID-AB2480080214>3.0.CO;2-L)
68. Mabelis A. 1984 Aggression in wood ants (*Formica polyctena* Foerst., Hymenoptera, Formicidae). *Aggress. Behav.* **10**, 47–53. (doi:10.1002/1098-2337(1984)10:1%3C47::AID-AB2480100107%3E3.0.CO;2-N)
69. Sorvari J, Hakkarainen H. 2004 Habitat-related aggressive behaviour between neighbouring colonies of the polydomous wood ant *Formica aquilonia*. *Anim. Behav.* **67**, 151–153. (doi:10.1016/j.anbehav.2003.03.009)
70. Stockan JA, Robinson E. 2015 Wood ant ecology and conservation. In *Ecology biodiversity and conservation* (eds JA Stockan, JH Robinson), p. 304. Cambridge, UK: Cambridge University Press. (doi:10.1017/CB09781107261402.014)
71. Driessen GJJ, Van Raalte At, De Bruyn GJ. 1984 Cannibalism in the red wood ant, *Formica polyctena* (Hymenoptera: formicidae). *Oecologia* **63**, 13–22. (doi:10.1007/BF00379779)
72. Borowiec ML, Cover SP, Rabeling C. 2021 The evolution of social parasitism in *Formica* ants revealed by a global phylogeny. *Proc. Natl Acad. Sci. USA* **118**, e2026029118. (doi:10.1073/pnas.2026029118)
73. Longino JT. 1989 Geographic variation and community structure in an ant–plant mutualism: *Azteca* and *Cecropia* in Costa Rica. *Biotropica* **21**, 126–132. (doi:10.2307/2388703)
74. Longino JT. 2020 *Azteca*. In *Encyclopedia of social insects* (ed. C Starr), pp. 88–92. Cham, Switzerland: Springer. (doi:10.1007/978-3-319-90306-4_13-1)
75. Yu DW, Davidson DW. 1997 Experimental studies of species-specificity in *Cecropia*–ant relationships. *Ecol. Monogr.* **67**, 273–294. (doi:10.1890/0012-9615(1997)067[0273:ESOSS]2.0.CO;2)
76. Sakata H, Katayama N. 2001 Ant defence system: a mechanism organizing individual responses into efficient collective behavior. *Ecol. Res.* **16**, 395–403. (doi:10.1046/j.1440-1703.2001.00404.x)
77. Gotwald WH. 1995 *Army ants: the biology of social predation*. Ithaca, NY: Cornell University Press.
78. Hoenle PO, Blüthgen N, Brückner A, Kronauer DJC, Fiala B, Donoso DA, Smith MA, Ospina Jara B, von Beeren C. 2019 Species-level predation network uncovers high prey specificity in a neotropical army ant community. *Mol. Ecol.* **28**, 2423–2440. (doi:10.1111/mec.15078)
79. Dejean A, Orivel J, Cerdá X, Azémér F, Corbara B, Touchard A. 2025 Foraging by predatory ants: a review. *Insect Sci* **32**, 1096–1118. (doi:10.1111/1744-7917.13461)
80. Chesson PL, Warner RR. 1981 Environmental variability promotes coexistence in lottery competitive systems. *Am. Nat.* **117**, 923–943. (doi:10.1086/283778)
81. Berghoff SM, Winter T, Gadau J, Linsenmair KE, Maschwitz U. 2003 Sociobiology of hypogaecic army ants: characterization of two sympatric *Dorylus* species on Borneo and their colony conflicts. *Insectes Sociaux* **50**, 139–147. (doi:10.1007/s00040-003-0642-z)
82. Sears ALW, Chesson P. 2007 New methods for quantifying the spatial storage effect: an illustration with desert annuals. *Ecology* **88**, 2240–2247. (doi:10.1890/06-0645.1)
83. Barden P, Grimaldi DA. 2016 Adaptive radiation in socially advanced stem-group ants from the Cretaceous. *Curr. Biol.* **26**, 515–521. (doi:10.1016/j.cub.2015.12.060)
84. Calcaterra LA, Livore JP, Delgado A, Briano JA. 2008 Ecological dominance of the red imported fire ant, *Solenopsis invicta*, in its native range. *Oecologia* **156**, 411–421. (doi:10.1007/s00442-008-0997-y)

85. Andersen AN, Blum MS, Jones TM. 1991 Venom alkaloids in *Monomorium 'rothsteini'* Forel repel other ants: is this the secret to success by *Monomorium* in Australian ant communities? *Oecologia* **88**, 157–160. (doi:10.1007/BF00320805)
86. Champer J, Schlenoff D. 2024 Battles between ants (Hymenoptera: Formicidae): a review. *J. Insect Sci.* **24**, 25. (doi:10.1093/jisesa/ieae064)
87. de Oliveira GV, Corrêa MM, Góes IMA, Machado AFP, de Sá-Neto RJ, Delabie JHC. 2015 Interactions between *Cecropia* (Urticaceae) and ants (Hymenoptera: Formicidae) along a longitudinal east-west transect in the Brazilian Northeast. *Annales de La Société Entomologique de France* **51**, 153–160. (doi:10.1080/00379271.2015.1061231)
88. Savolainen R, Vepsäläinen K, Wuorenrinne H. 1989 Ant assemblages in the taiga biome: testing the role of territorial wood ants. *Oecologia* **81**, 481–486. (doi:10.1007/BF00378955)
89. Vepsäläinen K, Pisarski B. 1982 Assembly of island ant communities. *Ann. Zool. Fenn.* **19**, 327–335.
90. Yéo K, Molet M, Peeters C. 2006 When David and Goliath share a home: compound nesting of *Pyramica* and *Platythyrea* ants. *Insectes Soc.* **53**, 435–438. (doi:10.1007/s00040-005-0890-9)
91. Parmentier T, Yéo K, Dekoninck W, Wenseleers T. 2017 An apparent mutualism between afro-tropical ant species sharing the same nest. *Behav. Ecol. Sociobiol.* **71**, 46. (doi:10.1007/s00265-017-2274-8)
92. Vida T, Calamari ZT, Barden P. 2025 Post K-Pg rise in ant and termite prevalence underlies convergent dietary specialization in mammals. *Evolution* **79**, 2315–2324. (doi:10.1093/evolut/qpaf121)
93. Ule E. 1901 Ameisengärten im Amazonasgebiet: Botanische Jahrbücher für Systematik. *Pflanzenesch. Pflanzengeogr.* **68**, 45–52.
94. Orivel J, Leroy C. 2011 The diversity and ecology of ant gardens (Hymenoptera: Formicidae; Spermatophyta: Angiospermae). *Myrmecol. News* **14**, 73–85. (doi:10.25849/myrmecol.news_014:073)
95. Chomicki G, Renner SS. 2016 Obligate plant farming by a specialized ant. *Nat. Plants* **2**, 16181. (doi:10.1038/nplants.2016.181)
96. Suarez AV, Goodisman MAD. 2021 Non-kin cooperation in ants. *Front. Ecol. Evol.* **9**. (doi:10.3389/fevo.2021.736757)
97. Kleinfeldt SE. 1986 Ant-gardens: mutual exploitation. In *Insects and the plant surface* (eds B Juniper, TRE Southwood), pp. 283–291. Oxford, UK: Edward Arnold.
98. Forel A. 1898 La parabiose chez les fourmis. *Bull. Société Vaudoise Sci. Nat.* **34**, 380–384.
99. Wheeler WM. 1921 A new case of parabiosis and the 'ant gardens' of British Guiana. *Ecology* **2**, 89–103. (doi:10.2307/1928921)
100. Davidson DW, Epstein WW. 1989 Epiphytic associations with ants. In *Ecological studies vascular plants as epiphytes* (ed. U Lüttge), pp. 200–233. Berlin, Germany: Springer. (doi:10.1007/978-3-642-74465-5_8)
101. Schmit-Neuerburg V, Blüthgen N. 2007 Ant-garden epiphytes are protected against drought in a Venezuelan lowland rain forest. *Ecotropica* **13**, 93–100.
102. Vantaux A, Dejean A, Dor A, Orivel J. 2007 Parasitism versus mutualism in the ant-garden parabiosis between *Camponotus femoratus* and *Crematogaster levior*. *Insectes Soc.* **54**, 95–99. (doi:10.1007/s00040-007-0914-0)
103. Vicente RE, Dáttilo W, Izzo TJ. 2014 Differential recruitment of *Camponotus femoratus* (Fabricius) ants in response to ant garden herbivory. *Neotrop. Entomol.* **43**, 519–525. (doi:10.1007/s13744-014-0245-6)
104. Leal LC, Jacovac CC, Bobrowiec PED, Camargo JLC, Peixoto PEC. 2017 The role of parabiotic ants and environment on epiphyte composition and protection in ant gardens. *Sociobiology* **64**, 276–283. (doi:10.13102/sociobiology.v64i3.1219)
105. Corbara B, Dejean A, Orivel J. 1999 Les jardins de fourmis, une association plantes-fourmis originale. *L'Année Biol.* **38**, 73–89. (doi:10.1016/S0003-5017(99)80027-0)
106. Prebus M, Rabeling C. 2025 Phylogenomics resolve the systematics and biogeography of the ant tribe myrmicini and tribal relationships within the hyperdiverse ant subfamily myrmicinae. *Syst. Biol.* **74**, 526–544. (doi:10.1093/sysbio/syaf022)
107. Blaimer BB. 2012 Acrobat ants go global-origin, evolution and systematics of the genus *Crematogaster* (Hymenoptera: Formicidae). *Mol. Phylogenetics Evol.* **65**, 421–436. (doi:10.1016/j.ympev.2012.06.028)
108. Lucky A, Trautwein MD, Guénard BS, Weiser MD, Dunn RR. 2013 Tracing the rise of ants: out of the ground. *PLoS One* **8**, e84012. (doi:10.1371/journal.pone.0084012)
109. Economo EP, Klimov P, Sarnat EM, Guénard B, Weiser MD, Lecroq B, Knowles LL. 2015 Global phylogenetic structure of the hyperdiverse ant genus *Pheidole* reveals the repeated evolution of macroecological patterns. *Proc. R. Soc. B* **282**, 20141416. (doi:10.1098/rspb.2014.1416)
110. Blaimer BB, Ward PS, Schultz TR, Fisher BL, Brady SG. 2018 Paleotropical diversification dominates the evolution of the hyperdiverse ant tribe *Crematogastrini* (Hymenoptera: Formicidae). *Insect Syst. Divers.* **2**, 1–14. (doi:10.1093/isd/ixy013)
111. Mehdiabadi NJ, Mueller UG, Brady SG, Himler AG, Schultz TR. 2012 Symbiont fidelity and the origin of species in fungus-growing ants. *Nat. Commun.* **3**, 840. (doi:10.1038/ncomms1844)
112. Schultz TR et al. 2024 The coevolution of fungus-ant agriculture. *Science* **386**, 105–110. (doi:10.1126/science.adn7179)
113. Ward PS, Brady SG, Fisher BL, Schultz TR. 2014 The evolution of myrmicine ants: phylogeny and biogeography of a hyperdiverse ant clade (Hymenoptera: Formicidae). *Syst. Entomol.* **40**, 61–81. (doi:10.1111/syen.12090)
114. Prebus M. 2017 Insights into the evolution, biogeography and natural history of the acorn ants, genus *Temnothorax* Mayr (Hymenoptera: Formicidae). *BMC Evol. Biol.* **17**, 250. (doi:10.1186/s12862-017-1095-8)
115. Burchill AT, Moreau CS. 2016 Colony size evolution in ants: macroevolutionary trends. *Insectes Soc.* **63**, 291–298. (doi:10.1007/s00040-016-0465-3)
116. Vizuela J et al. 2025 Adaptive radiation and social evolution of the ants. *Cell* **188**, 4828–4848. (doi:10.1016/j.cell.2025.05.030)
117. Boulay R, Arnan X, Cerdá X, Retana J. 2014 The ecological benefits of larger colony size may promote polygyny in ants. *J. Evol. Biol.* **27**, 2856–2863. (doi:10.1111/jeb.12515)

118. Matte A, LeBoeuf AC. 2025 Innovation in ant larval feeding facilitated queen-worker divergence and social complexity. *Proc. Natl Acad. Sci. USA* **122**, e2413742122. (doi:10.1073/pnas.2413742122)
119. Booher DB *et al.* 2021 Functional innovation promotes diversification of form in the evolution of an ultrafast trap-jaw mechanism in ants. *PLoS Biol.* **19**, e3001031. (doi:10.1371/journal.pbio.3001031)
120. Bokma F. 2009 Problems detecting density-dependent diversification on phylogenies. *Proc. R. Soc. B* **276**, 993–994. (doi:10.1098/rspb.2008.1249)
121. Etienne RS, Haegeman B, Stadler T, Aze T, Pearson PN, Purvis A, Phillimore AB. 2012 Diversity-dependence brings molecular phylogenies closer to agreement with the fossil record. *Proc. R. Soc. B* **279**, 1300–1309. (doi:10.1098/rspb.2011.1439)
122. Burbrink FT, Pyron RA. 2010 How does ecological opportunity influence rates of speciation, extinction, and morphological diversification in New World ratsnakes (tribe *Lampropeltini*)? *Evolution* **64**, 934–943. (doi:10.1111/j.1558-5646.2009.00888.x)
123. Barden P. 2017 Fossil ants (Hymenoptera: Formicidae): ancient diversity and the rise of modern lineages. *Myrmecol. News* **24**, 1–30. (doi:10.25849/myrmecol.news_024:001)
124. Jouault C, Swain A, Sosiak C. 2026 Diversity-dependent effects likely influenced the diversification of species-rich crown ant subfamilies during the Cenozoic [Data set]. Zenodo. (doi:10.5281/zenodo.18838407)
125. Jouault C, Swain A, Sosiak C. 2026 Supplementary material from: Diversity-dependent effects probably influenced the diversification of species-rich crown ant subfamilies during the Cenozoic. Figshare. (doi:10.6084/m9.figshare.c.8460429)